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Impact of erosion and accretion on the distribution of enterococci in beach sands

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ABSTRACT

Bacterial pathogens in coastal sediments may pose a health risk to users of beaches. Although recent work shows that beach sands harbor both indicator bacteria and potential pathogens, it is neither known how deep within beach sands the organisms may persist nor if they may be exposed during natural physical processes. In this study, sand cores of approximately 100 cm depth were collected at three sites across the beach face in Kitty Hawk, North Carolina, before, during, and after large waves from an offshore hurricane. The presence of DNA from the fecal indicator bacterium *Enterococci* was detected in subsamples at different depths within the cores by PCR amplification. Erosion and accretion of beach sand at the three sites were also determined for each sampling day. The results indicate that ocean beach sands with persisting enterococci signals could be exposed and redistributed when wind, waves, and currents cause beach erosion or accretion.

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1. Introduction

Beach water quality is monitored to reduce the risk of recreational water-borne illness. The abundances of the bacteria Escherichia coli (freshwater) or enterococci (marine) are used to assess whether water conditions are suitable for swimming. The presence of these indicator bacteria at densities higher than the EPA recommended levels has been linked to swimmer illnesses in epidemiological studies (Haile et al., 1999; Wade et al., 2003; Sinigalliano et al., 2010). Indicator bacteria do not cause disease, but are used to monitor beach water quality because they are associated with fecal material, and their occurrence in the environment was thought to be transient, indicating relatively recent contamination with raw or treated sewage or animal wastes. However, indicator bacteria and potential pathogens also exist in beach sands of both freshwater and marine environments (Ghinsberg et al., 1995; Obiri-Danso and Jones, 2000; Desmarais et al., 2002; Whitman and Nevers, 2003; Beversdorf et al., 2007; Yamahara et al., 2007; Hartz et al., 2008), where cell numbers can be substantially higher than in nearby waters (Burton et al., 1987; Doyle et al., 1992; Irvine and Pettibone, 1993; Oshiro and Fujioka, 1995; Whitman and Nevers, 2003; Yamahara et al., 2007). The origins of these organisms in sand often are unclear, but likely contain contributions from both exogenous (from sewage or

runoff) and indigenous sources (Ferguson et al., 2005). The presence, persistence, and regrowth of indicator bacteria in sediments suggest that beach sands may serve as a reservoir for these microbes, and that they may have the ability to impact water quality monitoring (Byappanahalli et al., 2003; Boehm and Weisberg, 2005; Beversdorf et al., 2007; Yamahara et al., 2007; Hartz et al., 2008; Halliday and Gast, 2011).

Studies regarding the movement of indicator bacteria between the sand and ocean have focused primarily on the effects of tides and rain (Boehm and Weisberg, 2005; Beversdorf et al., 2007; Yamahara et al., 2007; Zehms et al., 2008). These studies have observed increases in the water column levels of indicator organisms under high tide and rain events. Furthermore, a reduction in fecal indicator bacteria in sand approximately equal to the increase observed in the water column was reported (Yamahara et al., 2007). Even pedestrian traffic may influence the distribution of bacterial pollution at recreational beaches by translocation of bacterial-sized particles deposited on the sand surface (Bonilla et al., 2007). Several studies have examined the correlation between increased fecal indicator bacteria in sand and water within the swash zone (Zehms et al., 2008), and the potential for stream bank and bed resuspension to contribute to indicator presence in stream and lake water (Byappanahalli et al., 2003; Jamieson et al., 2005). The results support the hypothesis that sand and sediment disruption could contribute to the release of fecal indicator bacteria into water, but it is not known how changes in beach topology affect indicator distribution within the sand.

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Fecal indicator bacteria and pathogens from sewage spills or runoff introduced into the beach and adjacent shallow waters may pass into the subsurface sands. While many bacteria either will die or be grazed by protistan consumers, some may persist either as free particles or as particles incorporated into biofilms on grains of sand or small pebbles (Jeng et al., 2005), and may even regrow in this environment (Desmarais et al., 2002; Hartz et al., 2008; Yamahara et al., 2009). Waves can cause significant changes in the beach, through both accretion and erosion, and sediment that had been in contact with sewage could be buried and re-exposed. If pathogens were also present, human health hazards could be created by direct contact with contaminated sand and possibly by exposure to pathogens carried by spray from breaking waves in the surf. Eroded sediments from the beach can be transported long distances along the shoreline by surfzone currents (Smith and Largier, 1995; Ciavola et al., 1998; Grant et al., 2005; Feddersen, 2007), potentially distributing organisms as well.

To examine how physical forces may impact the distribution of fecal indicator bacteria, and potential pathogens, on beaches, the effects of sand erosion and accretion on the distribution of enterococci in ocean beaches were investigated. Here, it is shown that beach sands that may contain persistent populations of enterococci were exposed and reworked when large waves from an offshore hurricane caused erosion and accretion, potentially resulting in release of the organisms to the water column.

2. Methods

2.1. Location and sand core sampling

Sand cores were collected on a beach near Kitty Hawk (Byrd St. public access, Fig. 1), on the Outer Banks of North Carolina, between 11 and 15 September, 2006 as Hurricane Florence moved northward between Bermuda and the US coast (11 September, 2006). This site was selected because it is a well-attended public-access beach with numerous houses close to the shore (Fig. 1A), and thus the presence of fecal indicator bacteria was likely. There also were nearby environmental observations, and relatively accurate local wave forecasts that allowed determination of when beach change was likely to occur.

Winds and waves were measured at the Army Corps of Engineers Field Research Facility located about 10 km north of the study site (http://www.frf.usace.army.mil/). Wind speeds (Fig. 2A) and directions collected about 19 m above the water surface with an RM Young marine anemometer at the end of a pier (7 m water depth) were vector averaged and rotated so winds from true north have a direction of 0° (shore normal is 70°). Significant wave heights (defined as four times the standard deviation of the sea-surface elevation fluctuations, Fig. 2B) and directions were estimated from bottom-pressure fluctuations observed with an array of 15 bottom-mounted Setra gages just offshore of the pier in 8 m water depth using linear theory and a maximum likelihood estimator. Beach erosion and accretion were determined from walking surveys conducted on each of the sampling days by the Army Corps of Engineers using a differential Global Positioning System. The system has an accuracy of 3 cm in both the horizontal and vertical dimensions (personal communication, Mike Forte, US Army Corps of Engineers, Duck, NC).

Approximately 100 cm long sediment cores were collected in duplicate at three locations along a cross-shore transect between the high- and mid-tide regions (Fig. 1B). Cores were collected using new, clean acrylic tubes that were pounded into the beach surface by hand. After recovery from the beach, core ends were trimmed to be even with the sand surface and capped using new,



Fig. 1. Sampling site. (A) Google Earth image of the Byrd St. public access and the surrounding neighborhood and beach near Kitty Hawk, NC. Samples were taken along the red line, latitude=36.0983°N, longitude=75.7093°W. (B) Photograph of beach sites where cores were taken on September 11, 13, and 15, 2006.

clean plastic caps that were taped in place. Cores were collected on September 11, 13, and 15, and kept at 4 °C until sub-sampled on September 21, 2006 at the laboratory in Woods Hole, MA.

There were no on-site analysis facilities, and samples were preserved as best possible. The 7-day delay prior to processing would be inappropriate for culturing organisms, assessing community structure and diversity, or conducting quantitative PCR assays because there is the potential for loss of nucleic acids. For determining the presence or absence of a particular bacterial genus, the loss of target might result in a sample with an initially low signal to become negative in assays. However, DNA signals can persist for fairly long times in the environment. For example, PCR signals persisted for 28 days when seawater was spiked with enterococci, although the recovery of viable cells fell below the mEI agar plate detection level after five days (Walters et al., 2009). Seawater spiked with free DNA from enterococci vielded a PCR signal for 18 days (Walters et al., 2009), suggesting that whole cells in the dead or viable but not cultivable (VBNC) state may help preserve nucleic acids in the environment. Thus, it is expected that the results here are a conservative, but legitimate,



Fig. 2. (A) Wind speed and (B) significant wave height versus time from before, during, and after the sampling period. Waves were near shore normal throughout the study period, whereas winds were from north of shore normal until September 13, when wind directions became near shore normal then south of shore normal until September 15th (not shown).

assessment of the distribution of enterococci in the beach environment sampled.

Samples were extracted from the cores by drilling a 4 cm diameter hole through one side of the plastic core tube and inserting a 15 ml sterile tube through the hole to recover a small sand plug. Plugs from the same position in the duplicate cores were combined for extraction and amplification. All of the cores were sampled within 3-5 cm of the sand surface, and most also were sampled deeper in the core (between 14 and 70 cm below the surface). To ensure that subsequent samples from the same site were extracted at approximately the same vertical location (within 5 cm) as those in the first core (September 11), the depth of sampling was adjusted to account for erosion and accretion. For example, at site B there was 9 cm of erosion between September 11 and 13, and thus cores on September 11 were sampled at the surface, at 14 cm below the surface (corresponding to the surface of the core taken on September 13), and at 57 cm below the surface (corresponding to the September 13 sample 44 cm below the surface). An exception to sampling at approximately the same vertical location was made for the deeper samples at site A, which were taken from the sulfidic layer that was present between 22 and 25 cm below the surface, even as the site accreted. Subsamples of 0.25 g were recovered from the combined duplicate samples and were extracted using the MoBio PowerSoilTM DNA Kit following the manufacturer's instructions.

2.2. Detection of indicator organisms

A total of 23 samples were analyzed for DNA using PCR detection methodology. PCR was used to amplify the *Enterococcus* genus 23S ribosomal RNA gene using the ECST784F and ENC854R primers (Ludwig et al., 1994), and to amplify the *E. faecalis* 16S ribosomal RNA gene using the 72F and 210R primers (Sedgley et al., 2005). *E. faecalis* genomic DNA was purchased from ATCC[®] (700802) and used as a positive control for both primer sets. No-template DNA negative controls were included in every set of PCR reactions to monitor for contamination.

The 50 μ l PCR reaction mixture contained 1 \times Taq polymerase buffer, 100 ng each primer, 2.5 mM MgCl₂, 0.2 mM deoxynucleotides, 1.25 U GoTaq (Promega), 1 μ l of template DNA, and water. PCR conditions for ECST784F/ENC854R were 1 cycle of 95 °C for 2 min, followed by 40 cycles of 94 °C for 10 s, 52 °C for 30 s, 72 °C



Fig. 3. Test for detection limit of the PCR method. (A) *E. faecalis* genomic DNA dilution alone. (B) *E. faecalis* genomic DNA dilution plus 1:10 sand extract dilution. (C) *E. faecalis* genomic DNA dilution plus 1:100 sand extract dilution. (D) *E. faecalis* genomic DNA dilution plus 1:100 sand extract dilution. M=Promega 1 Kb ladder, 1=100 ng genomic DNA, 2=10 ng genomic DNA, 3=1 ng genomic DNA, 4=100 pg genomic DNA, 5=10 pg genomic DNA, 6=1 pg genomic DNA, 7=100 fg genomic DNA, 8=10 fg genomic DNA, (-)=n0 DNA negative control.

for 10 s, followed by 4 °C hold. The conditions were the same for primers 72F/210R, except that the annealing temperature was 62 °C. An ABI 2700 Thermal Cycler was used.

The detection limit of the PCR method for these samples was 1 pg of genomic DNA, roughly equivalent to between 28 and 280 cells (Fig. 3), determined using a 1:10 dilution series of *E. faecalis* genomic DNA (starting amount 100 ng). Extracted sand core DNA that was negative for the presence of enterococci using the genus primers was added to each reaction in the dilution series. Amplification inhibition was observed for reactions spiked with 1 μ l of undiluted sand core DNA (not shown, detection limit 10 ng), but dilution of the extract 1:10, 1:100, or 1:1000 mitigated this effect (Fig. 3). To reduce amplification inhibition, all DNA extracts were diluted 1:10 before amplification to test for the presence or absence of products for *Enterococcus* genus and *E. faecalis*.

3. Results and discussion

3.1. Beach dynamics and indicator presence

The offshore hurricane produced relatively light winds at the shoreline (Fig. 2A) that likely did not impact the beach topology as much as the large waves (Fig. 2B). On September 12, the waves and storm surge at high tide exceeded the top of the dunes at the public access parking lot area near the sampling site. At the high-tide site (A in Fig. 1B and Fig. 4) beach sediments accreted a total of 18 cm throughout the sampling period, whereas at sites B and C there was 8–9 cm of erosion during the storm, followed by 11–24 cm of accretion as the storm abated (Fig. 4).

None of the samples were positive by PCR for *E. faecalis*, but many produced a positive signal using the *Enterococcus* genus primers (Fig. 4). Surface samples at the high tide site (A, Fig. 4) always had a positive signal for enterococci, as were the samples collected deeper in the core. These results suggest the accreted sands could have contained enterococci.

The center site (B, Fig. 4) yielded enterococcal signals from the surface sands on the first day of sampling and in accreted surface sands at the end of the storm. This site was the only one that gave negative results for enterococcal products in subsurface samples at the start of the sampling. These negatives persisted through the second day of sampling as the site eroded, but they became



Fig. 4. Total enterococci PCR results from sand cores relative to overall beach change. Cores are labeled 1 (September 11), 2 (September 13), and 3 (September 15) at each site A, B, and C. '+' sign indicates core samples where enterococci were detected by PCR. '-' sign indicates core samples that were negative for enterococci PCR products. Elevations are cm relative to the North American Vertical Datum of 1988 (NAVD88, approximately 10 cm above Mean Sea Level). The arrows above each set of cores indicate erosion (downwards) and accretion (upwards), with the corresponding amounts of sand level change (cm) listed next to the arrows.

positive as sand accreted at the end of the storm, suggesting that deposited sands played a role in the redistribution of enterococcal signals.

At the mid-tide site (C, Fig. 4), the surface samples on September 11 were negative for enterococcal product, but subsurface samples were positive. Erosion at this site may have allowed the exposure of these subsurface sands, which became the enterococci positive surface sand samples on September 13. Accreted surface sand and subsurface sands both gave positive signals for enterococci on the final day of sampling.

Site C (mid-tide) was the only site that showed potential reworking of the sands as far as 60 cm below the surface. On the first day of sampling, the 60 cm-deep samples at site C had layers of fine sand, and gave positive PCR results for enterococci, but in the subsequent samples at that depth much coarser sands that were negative for enterococci products replaced the fine-sand layers. The change in enterococci presence at this sample location could be related to the grain size, as discussed below. However, the change in grain type also suggests that sands were moved to and from this location, and thus that enterococcal signals in the initial sample may have been transported to other sites. The erosion and subsequent cross-shore transport of enterococcicontaining sediments from the middle (B) and mid-tide (C) sites may have contributed to the enterococcal signals in accreted sands at all three sites.

The presence of enterococcal signals from 25–70 cm deep in the cores suggests there is potential for these indicator organisms to persist or regrow at significant depth in the beach environment. Natural populations of enterococci grow in the marine environment (Hartz et al., 2008; Yamahara et al., 2009), and can confound indicator studies. Thus, the signals present at depth, and even at the surface, in the samples reported here might be owing to indigenous strains of enterococci rather than the persistence of ones introduced by beach contamination. However, fecal enterococci can persist in beach sands longer than in the water column (Hartz et al., 2008), so it is possible that subsurface sediments could harbor fecal bacteria and pathogens.

The presence and abundance of fecal indicator bacteria can be influenced by the sand grain size within the beach, with more indicator bacteria present in fine sand with a uniform distribution (Skalbeck et al., 2010). Although grain sizes were not measured, visual inspections of the cores showed similar medium to coarsegrained sands (0.25–1.00 mm diameter) at all sample locations except the deep samples at site A and the initial (September 11) deep sample at site C. Thus, while the presence of an enterococcal signal at the deepest location at site C might be related to the fine grain size present prior to the storm, medium-coarse grained sands from other core samples also often were positive for enterococci. Therefore, it seems likely that the differences in enterococci detection were not owing to the differences in grain size only.

The PCR method detects the presence of nucleic acids, but not necessarily viable cells. Therefore, the results indicate that enterococcal genetic material was present in the samples. Positive results could have been caused by free DNA in the sample, as well as by viable, dead, or viable but not cultivable (VBNC) enterococcal cells. However, DNA signals from added cells have been shown to persist longer in environmental samples (28 days) than signals from added free DNA (18 days; Walters et al., 2009), suggesting that recovery of a DNA signal in the samples here likely indicates the presence of whole cells. Fecal enterococci have been found to persist in beach sands longer than in the water column (Hartz et al., 2008), so their presence in these samples would not be unprecedented. The extent to which human pathogens, such as Giardia or Staphylococcus, are able to persist within beach sands is unknown, although some of these pathogens have been detected (Abdelzaher et al., 2010; Plano et al., 2011).

4. Conclusions

The results from this pilot project demonstrate that there is potential for redistribution of enterococci as beach sand erodes or accretes. Unlike traditional culture approaches, the PCR-based method is useful because it can detect viable but not cultivable (VBNC) cells in addition to live cells. Future work with both culture-dependent and culture-independent methods is necessary to determine the extent that beach reworking contributes to the live enterococcal signal in water and throughout the beach environment.

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