

Accumulation of Fe oxyhydroxides in the Peruvian oxygen deficient zone implies non-oxygen dependent Fe oxidation

Maija I. Heller^{a,1}, Phoebe J. Lam^{a,*}, James W. Moffett^b, Claire P. Till^{a,2},
Jong-Mi Lee^a, Brandy M. Toner^c, Matthew A. Marcus^d

^a Department of Ocean Sciences, University of California, Santa Cruz, CA, United States

^b Marine Environmental Biology, Department of Biological Sciences, University of Southern California, Los Angeles, CA 90089, United States

^c Department of Soil, Water, and Climate, University of Minnesota-Twin Cities, St. Paul, MN 55108, United States

^d Advanced Light Source, Lawrence Berkeley National Laboratory, Berkeley, CA 94720, United States

Received 13 September 2016; accepted in revised form 12 May 2017; available online 19 May 2017

Abstract

Oxygen minimum zones (OMZs) have been proposed to be an important source of dissolved iron (Fe) into the interior ocean. However, previous studies in OMZs have shown a sharp decrease in total dissolved Fe (dFe) and/or dissolved Fe (II) (dFe(II)) concentrations at the shelf-break, despite constant temperature, salinity and continued lack of oxygen across the shelf-break. The loss of both total dFe and dFe(II) suggests a conversion of the dFe to particulate form, but studies that have coupled the reduction-oxidation (redox) speciation of both dissolved and particulate phases have not previously been done. Here we have measured the redox speciation and concentrations of both dissolved and particulate forms of Fe in samples collected during the U.S. GEOTRACES Eastern tropical Pacific Zonal Transect (EPZT) cruise in 2013 (GP16). This complete data set allows us to assess possible mechanisms for loss of dFe. We observed an offshore loss of dFe(II) within the oxygen deficient zone (ODZ), where dissolved oxygen is undetectable, accompanied by an increase in total particulate Fe (pFe). Total pFe concentrations were highest in the upper ODZ. X-ray absorption spectroscopy revealed that the pFe maximum was primarily in the Fe(III) form as Fe(III) oxyhydroxides. The remarkable similarity in the distributions of total particulate iron and nitrite suggests a role for nitrite in the oxidation of dFe(II) to pFe(III). We present a conceptual model for the rapid redox cycling of Fe that occurs in ODZs, despite the absence of oxygen.

© 2017 Elsevier Ltd. All rights reserved.

Keywords: Oxygen minimum zone; Oxygen deficient zone; OMZ; ODZ; Fe redox cycling; Synchrotron XAS

1. INTRODUCTION

1.1. Fe in seawater

Iron (Fe) is the most important trace metal in marine biogeochemistry due to its role as a key limiting element to

primary productivity in many regions of the ocean (Boyd et al., 2007). Fe exists in two redox states in the environment, Fe(II) and Fe(III), with Fe(III) being the thermodynamically favored form in oxygenated seawater. However, dissolved Fe (III) (dFe(III)) becomes strongly hydrolyzed in seawater and forms Fe oxyhydroxides, which have a very low solubility.

* Corresponding author.

E-mail address: pjlam@ucsc.edu (P.J. Lam).

¹ Present address: Ifremer, Laboratoire des Cycles Géochimiques, CS 10070, F-29280, Plouzané, France.

² Present address: Chemistry Department, Humboldt State University, Arcata, CA, United States.

Dissolved Fe(II) (dFe(II)) is the more soluble form of Fe, but exists in ambient oxic seawater at extremely low concentrations because of rapid oxidation by its two main oxidants, oxygen (O_2) and hydrogen peroxide (H_2O_2) (Millero et al., 1987; Millero and Sotolongo, 1989). In surface oxygenated seawater, dFe(II) is primarily generated by photochemical processes (Johnson et al., 1994), and is viewed as a reactive intermediate in Fe cycling.

DFe(II) exists at higher concentrations in oxygen depleted marine sediments where it is produced by dissimilatory Fe reducers that use Fe(III) to oxidize organic matter (Noffke et al., 2012). If the water overlying these sediments is oxygenated, reduced dFe(II) that diffuses from sediments is quickly oxidized by O_2 or H_2O_2 to dFe(III), which forms colloidal and particulate Fe(III) (pFe(III)) species. However, if the overlying waters are suboxic, the main oxidants for Fe(II) are absent, and consequently the Fe(II) redox state is more stable and can exist at nanomolar levels in the dissolved phase in oxygen deficient regions (Kondo and Moffett, 2013; Vedamati et al., 2014).

So-called oxygen minimum zones (OMZs) in the ocean are defined as regions that show distinct low oxygen layers at intermediate depths (~100 to 1000 m) (Bopp et al., 2002; Stramma et al., 2008). Global OMZ volumes are predicted to increase with climate change, with some indications that this is already happening (Stramma et al., 2008). OMZs are generally found in waters that have been isolated from the atmosphere due to sluggish circulation, and thus have had their oxygen consumed by the oxidation of organic matter with no chance for renewal. Globally there are three major open ocean areas that exhibit this pattern—the eastern tropical Pacific, the eastern tropical Atlantic, and the Arabian Sea. Within these, the eastern tropical Pacific and Arabian Sea have “oxygen deficient zones” (ODZs), which are defined as areas in which dissolved oxygen is undetectable (Revsbech et al., 2009).

OMZs have been proposed to be a source of Fe to the interior ocean through an “open marine iron shuttle” mechanism (Scholz et al., 2014b), in which elevated dFe(II) that is reductively mobilized from shelf sediments is exported into the ocean interior within the ODZ (Lyons and Severmann, 2006; Raiswell and Canfield, 2012; Scholz et al., 2014b). Scholz et al. (2014a) argue that Fe release from continental margin sediments is most effective in a narrow redox window where neither oxygen nor sulfide is present. The implications for this are that future deoxygenation in the Peru upwelling area might not result in increased Fe availability if the water column becomes sulfidic, but partial deoxygenation may enhance the Fe supply to the ocean from weaker OMZs.

The open marine Fe shuttle idea of Fe transport to the interior was based on a study of excess Fe (particulate Fe in sediments in excess of an expected lithogenic component) in surface sediments. It hypothesizes that dFe(II) should be high throughout the ODZ, and be removed at the oxycline below and at the offshore edge of the ODZ through oxidation and precipitation as particulate Fe(III) (pFe(III)).

However, in transects offshore from the broad Peruvian shelf, water column studies found that both total dFe and dFe(II) dropped off quickly with distance from the coast.

Hong and Kester (1986) identified a chemical front about 30 km offshore in which both dFe(II) and total Fe dropped off sharply (from ~30 nM to ~1 nM), showing that most Fe appeared to be trapped on the shelf. Likewise, both Bruland et al. (2005) and Vedamati et al. (2014) confirmed the sharp decrease in total dFe and/or dFe(II) concentrations at the shelf-break, despite constant temperature, salinity and continued lack of oxygen across the shelf-break.

The loss of both total dFe and dFe(II) suggests a conversion of the dFe to particulate form. The best evidence for this was a study by Landing and Bruland (1987), who measured total dFe as well as reactive (leachable in acetic acid) and total pFe at several stations in the Eastern Tropical North Pacific. At one offshore station, they found that the fraction of reactive pFe increased slightly in the ODZ. They hypothesized that reduced dFe was scavenged to a reactive particulate form, potentially low crystallinity oxyhydroxides, but they did not make redox measurements of dissolved or particulate Fe. Similarly, Vedamati et al. (2014) concluded that large lateral gradients observed in dissolved Fe across the shelf-slope break reflected production of dFe(II) by benthic processes on the shelf and dFe(II) removal by “oxidative scavenging”, but they did not make measurements of the particulate phase.

Our perspective on redox processes in ODZs has been transformed by new measurements of dissolved oxygen enabled by STOX sensors (Revsbech et al., 2009; Tiano et al., 2014), which reveal that dissolved O_2 levels are less than 10 nM, much lower than the micromolar detection levels of conventional sensors or chemical assays. Tiano et al. (2014) argued that with detection limits this low, these waters are “functionally anoxic”. Indeed, it seems unlikely that oxygen could be poised at the low nanomolar level over hundreds of meters and is probably truly zero. Ulloa et al. (2012) went further and argued that ODZs with no STOX-detectable oxygen should be renamed Anoxic Marine Zones (AMZ). This suggests that we must look closely at linkages between Fe redox chemistry and the nitrogen cycle, since nitrate and nitrite are the only remaining plausible oxidants of Fe. Strong evidence for nitrate oxidation of Fe(II) derived from shelf sediments has been reported by Scholz et al. (2016), who also reported genomic evidence for the presence of microbes containing nitrate reductases that have been shown in other systems to oxidize Fe(II).

The U.S. GEOTRACES Eastern tropical Pacific Zonal Transect (EPZT) cruise in 2013 (GP16) provided an opportunity to study, for the first time, Fe redox chemistry in both the dissolved and particulate phases in the water column on a transect extending to the westernmost limits of the ODZ. It also enabled us to place our findings in the context of other observations, including the distribution of other redox-reactive elements and radium-228 (Ra-228), which, like Fe(II), is also derived from marine sediments. This paper provides important insight into how processes studied by Scholz and coworkers in near bottom waters on the continental margin (Scholz et al., 2016) can be extended up into the water column and west to the offshore ODZ, and shows that rapid redox cycling of iron occurs in ODZs, despite the absence of oxygen.

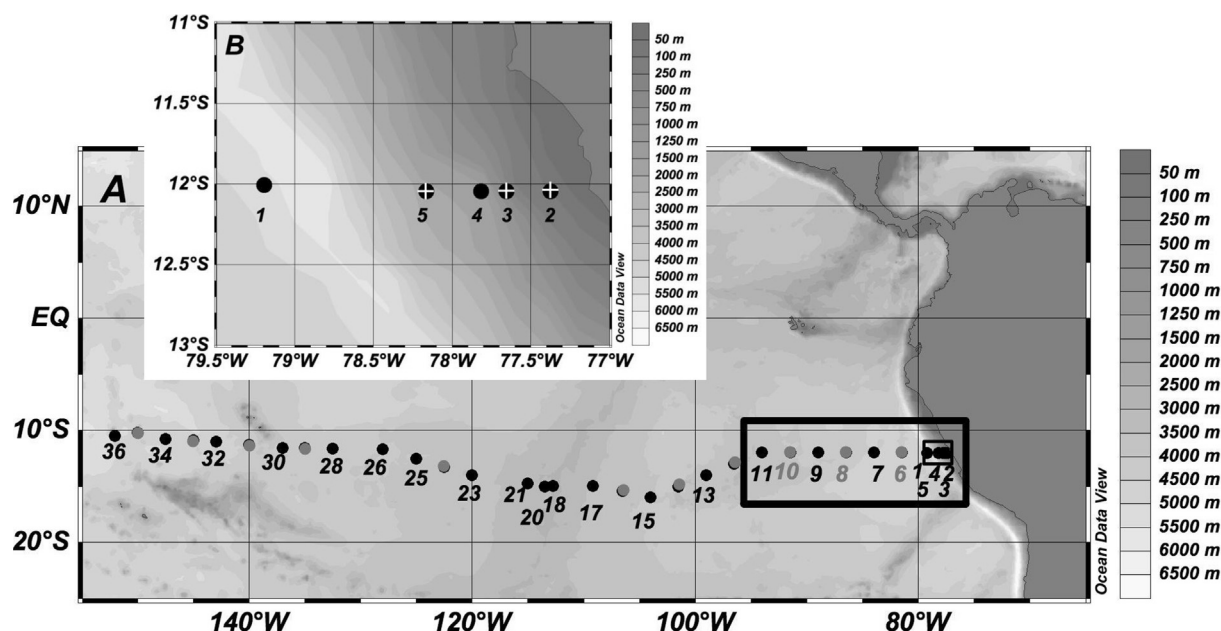


Fig. 1. (A) Cruise track of the Eastern Pacific Zonal Transect (EPZT—GP16) cruise. All full stations (full ocean depth) indicated in black dots; “demi” stations (to 1000 m only) in grey dots. The thick black box denotes the eastern portion (stations 1–11) of the cruise discussed in this paper. Hydrographic parameters (O_2 , NO_3^- , NO_2^-) and total dissolved Fe measured at all stations ($n = 11$); dissolved Fe(II) and total particulate Fe measured at full stations only ($n = 8$). (B) Greater bathymetric detail for Stations 1–5, outlined in thin black box in (A), showing stations (black dots with white crosses) where synchrotron analyses of pFe were done ($n = 3$). Note that station numbers increase offshore, except for station 1, which is out of sequence and is offshore of Station 5. Background shading and scale bar show bathymetry.

2. STUDY AREA AND SAMPLING STRATEGY

The U.S. GEOTRACES Eastern Pacific Zonal Transect cruise (GEOTRACES cruise GP16) took place on the RV Thomas G. Thompson (TT-303) from 25 October to 20 December 2013 (Fig. 1). The study area lies within the Humboldt Current system where southerly winds are known to cause an offshore Ekman transport and coastal upwelling. The Peru undercurrent, which is derived from the Equatorial Undercurrent, is in general known to be nutrient rich but low in oxygen (Kessler, 2006). The upwelling of this nutrient rich water is responsible for the high primary productivity at the Peruvian coast. Poor ventilation and the high oxygen demand are responsible for the OMZ in this region, which extends westward from the shelf from near surface to a depth of ~ 1000 m.

Three sampling systems were used to collect samples reported here: the 12-bottle Scripps Oceanographic Data Facility (ODF) rosette for non-contamination prone elements, the U.S. GEOTRACES sampling system of 24 Teflon-coated GO-FLO bottles (GTC Rosette) (Cutter and Bruland, 2012) for contamination prone elements (dFe(II) and dFe), and modified McLane WTS-LV in-situ pumps for particulate samples. High quality hydrographic parameters for oxygen, nitrite, nitrate measurements and shipboard data management were provided by the Scripps ODF for water from both rosettes. Nutrient analyses were performed on a Seal Analytical continuous flow Auto Analyzer 3. Nitrate was measured by first reducing it to nitrite using a cadmium reduction column. Nitrite was measured by standard colorimetric analysis of the red azo dye.

Detection limits for $NO_3^- + NO_2^-$, and NO_2^- were each $0.02 \mu M$. Dissolved oxygen analyses were performed using a whole-bottle modified Winkler titration method using an automated oxygen titrator. Detection limit was $0.90 \mu mol/kg$. Standard WOCE and CLIVAR protocols were used for all hydrographic measurements according to <http://www.go-ship.org/HydroMan.html>; further descriptions of analytical procedures, precisions, and calibration details are available at: <http://cchdo.ucsd.edu/cruise/325020131025>. Nitrite data from the ODF rosette cast(s) closest to the pump casts are used for comparisons to pump particles. Oxygen profiles are from the upcast of the SBE43, a polarographic membrane oxygen sensor, mounted on the ODF rosette, which was calibrated by Winkler titration noted above. The SBE43 typically has a detection limit of $1\text{--}2 \mu mol O_2/kg$. Full resolution data for each system will be available at BCO-DMO (<http://data.bco-dmo.org/jg/dir/BCO/GEOTRACES/EPZT/>).

2.1. Particulate sampling

Sample collection was via dual-flow in-situ battery-powered pumps (modified McLane WTS-LV) suspended from a non-metallic hydrowire (Hytrel-coated Vectran). In-situ pump operations on GP16 were the same as used for GA03, the U.S. GEOTRACES North Atlantic Zonal Transect, described in (Ohnemus and Lam, 2015). Briefly, pumps operated for 4 h at a programmed rate of 8 L/min in a dual-flowpath configuration that allowed for simultaneous collection of particles on two filter types: $51 \mu m$ polyester prefilters were upstream of paired $0.8 \mu m$

polyethersulfone (PES) filters (Supor800, 142 mm diameter, 126 mm active diameter) on one flowpath, and also upstream of paired quartz fiber filters (Whatman QMA) on the other flowpath. All filters were acid-leached prior to use (Cutter et al., 2014) and loaded in mini-MULVFS filter holders (Bishop et al., 2012). All sample handling was conducted in a self-built HEPA-filtered clean room bubble using trace-element handling techniques and acid-cleaned plastic equipment. Wedge-shaped subsamples of 51 μm pre-filters and paired Supor filters were cut using a ceramic rotary blade.

As soon as possible after the filter holders were brought on board, a small subsample (0.9 cm^2 , representing ~ 3 L of equivalent volume filtered) of the Supor filter was cut and transferred into a glove box with a N_2 atmosphere, sealed in a gas impermeable Mylar bag, and kept frozen to retard further oxidation of Fe in samples collected from the ODZ. This subsample was used for synchrotron analysis of pFe speciation.

The rest of the Supor filter was dried at room temperature on acid-leached polystyrene “eggcrate” grids in a HEPA-filtered flow hood, and subsequently stored in clear polyethylene clean room bags (KNF Flexpak). Samples from Stations 1–4 were lightly misted with Milli-Q water over gentle vacuum to reduce sea salt retention before drying; samples from Stations 5 and later were simply sucked down over gentle vacuum before drying. Comparisons of misted and unmisted samples from one station (Station 17) showed no significant difference in the concentrations of non-salt-dominated elements determined by ICP-MS.

Dipped blank filters were also deployed on each cast to allow seawater exposure to blank filters during deployment. These were complete filter sets that were processed identically to samples and thus functioned as process blanks (Bishop et al., 2012; Ohnemus and Lam, 2015). For Stations 1–21, dipped blank filter sets were loaded in mini-MULVFS filter holders that were not connected to the pump plumbing with a 0.2 μm Supor prefilter to exclude particles.

2.2. Seawater sampling

Dissolved seawater samples were collected using the U. S. GEOTRACES sampling system of 24 Teflon-coated GO-FLO bottles (GTC Rosette) (Cutter and Bruland, 2012). At each station, the bottles were deployed open and tripped on ascent at 3 m/min. GO-FLOs were sampled in the GEOTRACES trace metal clean sampling van that contained HEPA-filtered air. GO-FLO bottles were pressurized to ~ 0.4 atm with 0.2 μm -filtered nitrogen, in contrast to previous GEOTRACES cruises where compressed air was used. Redox sensitive parameters were sampled first. Samples for total dissolved Fe were filtered using an Acropak-200 Supor capsule filter (0.2 μm pore size made of polyethersulfone). Before use, the Acropak filter was filled with acidified filtered surface seawater (pH 2, Fisher TraceMetal™ grade HCl) and left overnight to soak. Before collecting any subsamples, at least 500 mL of seawater was passed through the filter. Sample bottles for dFe were rinsed three times with 10% sample volume before filling. Acropak filters were used for at most 3 casts before a new

filter was used, and they were stored empty in a refrigerator while not in use.

For Fe(II) determination, samples were collected downstream of the Acropak, or in some cases, downstream of 0.2 μm pore size Supor filters directly into polypropylene/polyethylene rubber-free syringes (Air-Tite, VWR) via luer fittings. They were stored in a bucket containing ice and analyzed in a clean van adjacent to the sampling van within 20 min.

3. ANALYTICAL METHODS

3.1. Dissolved Fe concentrations

Samples for dFe were filtered into LDPE bottles that had been cleaned in accordance with the GEOTRACES cookbook (<http://www.geotraces.org/images/stories/documents/intercalibration/Cookbook.pdf>). Samples were acidified at sea with the equivalent of 4 mL of 6 M quartz-distilled HCl (q-HCl) per liter of seawater to pH 1.7–1.8, and were left acidified for at least several months prior to analysis in the laboratory. Dissolved Fe was analyzed using an adaptation of Biller and Bruland (2012) as described in Parker et al. (2016). Briefly, this method involves UV irradiating the samples, pre-concentrating the metals of interest by loading seawater at pH 6.0 onto a column of Nobias-chelate PA-1 resin and eluting with 1 N q- HNO_3 , and analyzing the eluent on the Thermo-Element High Resolution XR Inductively Coupled Plasma Mass Spectrometer at UC Santa Cruz. Iron was analyzed in medium resolution in counting mode.

Two different blanks were determined and subtracted from sample values (Parker et al., 2016). Column blanks were analyzed each ICP-MS run by loading only a small volume (~ 1 mL) of low-metal seawater and otherwise pre-concentrating and analyzing on the ICP-MS as usual. The column blanks serve to alert us to any differences in the blank between columns and also account for the blank of the analytical process, except from the reagents in the samples. For that we also analyze reagent blanks each ICP-MS run by preconcentrating low-metal seawater twice with usual concentrations of reagents and twice with double the reagents, and subtracting the resulting concentrations. The column blanks were used to calculate the detection limit (defined as three times the standard deviation of the blank), which was 0.038 nM ($n = 34$) during analysis of these samples.

As a quality control, SAFe reference samples were analyzed in the same preconcentration and ICP-MS runs as the EPZT samples. Our analysis resulted in values for SAFe S of 0.095 ± 0.014 nmol/kg ($n = 12$), for SAFe D2 of 0.965 ± 0.089 nmol/kg ($n = 10$), and for SAFe D1 of 0.68 ± 0.07 ($n = 6$). These values are within one standard deviation of the May 2013 consensus values of 0.093 ± 0.008 , 0.933 ± 0.023 , and 0.67 ± 0.04 nmol/kg, respectively.

3.2. Dissolved Fe(II) concentrations

Dissolved Fe(II) concentrations were determined using an automated flow injection analysis system (FeLume II

Waterville Analytical) employing a luminol chemiluminescence based detection system (King et al., 1995). Luminol, prepared in 18.2 MΩ Milli-Q water and buffered to ~pH = 10.3 with ammonia (Optima™), reacts with an Fe(II)-containing solution, resulting in luminol oxidation with concurrent chemiluminescent emission (Rose and Waite, 2001; Croot and Laan, 2002). The FeLume was fitted with a standard quartz flow cell and a Hamamatsu HC135 photon counter configured with the following settings: flow rate: 2.5 mL per minute; photon counter integration time: 200 ms; load time: 20–40 s. The mixing and reaction occur in a spiral flow cell positioned in front of a photomultiplier tube. The sample and luminol reagent was continually mixed in the flow cell by omitting the injection valve (Rose and Waite, 2001; Hopkinson and Barbeau, 2007; Roy et al., 2008). Once the signal was in steady-state, the mean of the last 50 data points was used to determine the signal.

Fe(II) was quantified by standard additions of Fe(II) (typically 0.05, 0.1, 0.5 and 1 nM) to the samples of surface seawater (stored in the dark several days) or deep, oxygenated water with low initial signals. A 0.01 M Fe(II) stock solution was prepared with ferrous sulfate (Sigma) in 0.2 M HCl. A working Fe(II) standard solution (100 nM) was prepared daily. Detection limits were determined for aged surface samples where ferrous Fe was negligible based on a standard 3σ evaluation of the baseline signal (Moffett et al., 2007; Kondo and Moffett, 2013). This led to a detection limit of 14 pmol/L.

3.3. Particulate trace metal concentrations

Filter subsamples of the top Supor filter representing 6–12% of the filter area (about 25–50 L equivalent volume filtered) were digested for determination of total particulate trace metal concentrations in the 0.8–51 μm size fraction. Digestion followed the two-step method (Ohnemus et al., 2014; Ohnemus and Lam, 2015), in which the polyethersulfone (PES) filter is first completely dissolved in the highly oxidizing Piranha reagent, which is 3 parts concentrated (18.4 M, 98%) sulfuric acid to 1 part concentrated (30–32%) hydrogen peroxide (H₂O₂), and then lithogenic material is dissolved in a strong acid mixture containing 4 M each of HNO₃, HCl, and HF (Ohnemus et al., 2014). All reagents were Seastar Baseline-grade unless otherwise mentioned. All digestions and PFA vial cleaning were conducted in a HEPA-filtered venting fumehood at University of California, Santa Cruz (UCSC), within a HEPA-filtered room.

Slight modifications were made to minimize reagent usage in the Piranha step to accelerate the dry-down process. Filter subsamples were cut into small pieces using a ceramic blade before placing in 15 mL PFA vials (Saville) to ensure that the entire filter material was covered by the digestion solution. Initially, 800 μL concentrated H₂SO₄ was added to the vial containing the cut filter pieces, along with 20 μL of a rhodium (Rh) spike (final concentration 1 ppb) to monitor potential loss during the digestion process. The vials were heated at 150 °C for 1.5 h with the lids on loosely. The hotplate was then turned off, the lids taken off, and each vial was gently swirled. After ~5 min, 265 μL

concentrated H₂O₂ was incrementally added to each vial, generating the Piranha solution. After the first addition of H₂O₂, the hotplates were turned on again at 100 °C followed by a second 265 μL H₂O₂ addition. The temperature was then increased to 150 °C for 30 min, followed by a third addition of 265 μL H₂O₂. After each addition of H₂O₂, the lids were placed loosely atop the vials to allow a good reactivity of the Piranha solution while also preventing loss associated with spatter.

Depending on the particle load and composition on each filter, three H₂O₂ additions resulted in liquid ranging from colorless to dark brown, with occasionally undissolved filter pieces. In samples that were not colorless, additional smaller H₂O₂ volumes (50–150 μL) were added and the vials swirled until subsequent H₂O₂ additions no longer generated a fizzing solution and a clear solution was attained. Vial blanks (reagents only) and dipped filter blanks underwent the same procedures as the samples. After all solutions were clear, lids were removed, and vials and inverted lids were carefully dried, first at 200 °C, then 220 °C, and finally 235 °C for several hours until only a small droplet remained.

After the PES matrix was oxidized by the Piranha procedure, a mineral acid total digestion was performed on the remaining particulate material (Ohnemus et al., 2014; Ohnemus and Lam, 2015). Briefly, 2 mL of a freshly prepared mixture of HNO₃, HCl, and HF (4 M each) was refluxed for 4 h at 135 °C, then dried down for several hours at 110 °C.

The final remaining droplet was brought up in 2 mL 5% HNO₃ with 1 ppb In as an internal standard. The final digest solution was diluted gravimetrically 5x and 20x using the same 5% HNO₃/1 ppb In solution and analyzed by high-resolution Inductively Coupled Mass Spectrometry (HR-ICP-MS) using an Element-XR ICP-MS (Thermo Scientific) with SC-FAST (Elemental Scientific) sample introduction system and PC3 Peltier cooled quartz cyclonic spray chamber in the UCSC Plasma Analytical Facility. Instrument drift and matrix effects were corrected using the internal 1 ppb in standard and monitored using a mixed element run standard. Concentrations were determined using external standard curves of mixed trace elements standards. The median dipped blank (15.0 nmol Fe/filter) was subtracted from all samples. The detection limit (58.7 nmol Fe/filter) was defined as three times the standard deviation (19.6 nmol Fe/filter) of 45 dipped blank filters, which translates to an approximate concentration detection limit for pFe of 0.14 nM given typical volumes filtered. The resulting Fe data for the eastern third of the transect is discussed in this present manuscript. Other particulate trace metal data will be discussed in subsequent manuscripts.

The same digestion and analysis procedure was regularly applied to approximately 15 mg each of two certified reference materials: BCR-414, a freshwater plankton, and PACS-2, a coastal marine sediment. Mean ± standard deviation of recoveries for Fe were 106 ± 13% for BCR-414 and 83 ± 6% for PACS-2.

3.4. Particulate iron speciation

We conducted a series of synchrotron-based X-ray measurements of the 0.8–51 μm particulate size fraction col-

lected by in-situ filtration during 3 visits to the Advanced Light Source (ALS) microprobe beamline 10.3.2 in Berkeley, CA, USA (Marcus et al., 2004) in 2014 and 2015. Beamline 10.3.2 is a bending magnet microprobe beamline that uses a silicon (Si(111)) crystal monochromator to create a monochromatic beam with a 2–15 μm spot size. All measurements were made in fluorescence mode using a Canberra 7-element UltraLEGe solid-state Ge detector. For all measurements, the monochromator was calibrated to put the inflection point for Fe metal at 7110.75 eV.

All samples were mounted dry in a manner that limited sample exposure to oxygen during irradiation by X-ray beam (Zeng et al., 2013; Toner et al., 2016). Briefly, all sample preparation was conducted in a portable Captair pyramid glove bag under argon (Ar) atmosphere. Samples, which were stored under Ar or N_2 atmosphere at -20°C , were mounted and sealed between two layers of X-ray transparent, gas-impermeable Mylar film. The X-ray analysis was then performed at ambient atmosphere.

3.4.1. Chemical species mapping

Chemical species maps were created to determine the redox speciation of particulate Fe in our samples. X-ray Fluorescence (XRF) maps were collected at four energies around the Fe K-edge (Table 1). Typically, a randomly chosen $1000\ \mu\text{m} \times 250\ \mu\text{m}$ area was mapped using a beam spot size of $5\ \mu\text{m}$, a pixel size of $4\ \mu\text{m}$, and a dwell time of 100 ms. The mapping area represented an equivalent volume filtered of $\sim 8\ \text{mL}$. Differences in sample to detector distances or number of functioning detector elements between beamline visits were accounted for using a thin-film XRF standard (Micromatter™, Vancouver, Canada) for Fe ($43.9\ \mu\text{g Fe/cm}^2$), which was mapped at the same beamline settings as the samples.

The pre-edge (7100 eV) and post-edge (7296 eV) energies to determine the background and total FeK α counts, respectively, were identical for all beamline visits. The two intermediate energies used to distinguish between Fe species varied slightly between beamline visits (Table 1) as we optimized the analysis over the beamline visits with the help of an error minimization and analysis program available at the beamline (Marcus, 2010; Lam et al., 2012). We chose the following three Fe species as reference input spectra for calculating chemical species maps: pyrite to represent Fe sulfides, FeSO_4 to represent Fe(II)-containing species, and a hydrous ferric oxide (“HFO-Fe4-P4”) to represent Fe(III) species. The XANES spectra of both FeSO_4 and HFO-Fe4-P4 have very little structure post edge, and were chosen specifically to represent generic Fe(II)- and Fe(III)-containing species, respectively. The chemical species mapping calculation partitions total apparent Fe counts at 7296 eV into four channels: background counts from scat-

ter and non-Fe sources, Fe sulfides, Fe(II) species, and Fe (III) species. The proportion of Fe attributed to each Fe species (Fe sulfide, Fe(II), and Fe(III)) was determined as the ratio of counts in each Fe species channel to the sum of the three Fe species. Total pFe concentrations were calculated from background-corrected counts in the 7296 eV channel using the Fe standard to $\mu\text{g Fe/cm}^2$ and dividing by the equivalent volume filtered through the map area. Concentrations of pFe species were calculated from total pFe and the proportion of Fe attributed to each Fe species.

3.4.2. X-ray absorption spectroscopy

The attribution of Fe species from the chemical species mapping was verified by collecting multiple Fe K-edge extended $\mu\text{-X-ray}$ Absorption Near-Edge Structure ($\mu\text{-XANES}$) spectra for each sample. A reproducible glitch in the monochromator at 7263.8 eV was used as an internal energy calibration for each extended XANES spectrum. All Fe K-edge $\mu\text{-XANES}$ data were collected from 7010 to 7414 eV in quick-XAS mode ($\sim 45\ \text{s/sweep}$), with the finest energy resolution (0.5 eV) in the XANES region (7095–7140 eV) and near the monochromator glitch (7261–7271 eV), and coarser step sizes otherwise (1 eV between 7010–7095 eV and 7140–7261 eV, and 3 eV between 7271–7414 eV). Sweeps were repeated until signal to noise was acceptable: typically 20–30 sweeps for each spot. For each spectrum, we corrected for detector deadtime and energy drift, removed the pre-edge background, and normalized the spectrum based on the extended region XANES region (~ 7250 to $7350\ \text{eV}$). Normalized XANES spectra were fit using linear combination fitting of a library of extended Fe K-edge XANES spectra from 58 Fe reference minerals that represented a range of Fe(II) and Fe(III) containing oxides, silicates, carbonates, and sulfides (Table 2).

The identity of Fe(III) species in the ODZ was determined using Extended X-ray Absorption Fine Structure (EXAFS) spectroscopy. EXAFS data were collected in quick-XAS fluorescence mode ($\sim 1\ \text{min/sweep}$) from 7000 to 7800 eV (to $k = 13$). More than 200 sweeps were collected for each spot for adequate signal to noise in the EXAFS region. Sweeps were averaged, deadtime corrected, and pre-edges were removed. E_0 was determined as the energy halfway up the edge step, and was 7123 eV for both samples. A cubic spline fit between 7131 and 7765 eV ($k = 1.5\text{--}13$) was used to remove the background and extract the EXAFS. The k^3 -weighted EXAFS were fit by linear combination fitting of a library of EXAFS spectra of Fe(III) species (Table 2). The ten best fits were examined to assess the robustness of the species appearing in the fits.

There was no evidence of beam damage over the course of the multiple XANES or EXAFS sweeps. For all XANES and EXAFS, the incident beam spot size was adjusted to

Table 1
Energies used for chemical species mapping and Fe standard pixel densities (cps/ μm^2) for the three beamline visits.

Beamline visit (yyyy–mm)	E1 (eV)	E2 (eV)	E3 (eV)	E4 (eV)	Fe standard (cps/ μm^2)
2014–04	7100	7116.5	7123	7296	10,980
2014–09	7100	7117.3	7123 or 7129.6 (GT2443)	7296	12,363
2015–10	7100	7117.5	7124.5	7296	7769

Table 2

Model Fe sulfide, Fe(II)-containing, and Fe(III) model compounds used in linear combination fits for XANES and EXAFS spectra (indicated by “x”). Abbreviations are those suggested by Whitney and Evans (2010) where possible, or for Source Clays, abbreviations used by the Clays and Clay Minerals society (<http://www.clays.org/SOURCE%20CLAYS/SCdata.html>).

Mineral	Mineralogical classification	Abbrev.#	Chemical formula	XANES library	EXAFS library
<i>Fe sulfides</i>					
Pyrite	Sulfide	Py	Fe^{2+}S_2	x	x
Pyrrhotite	Sulfide	Po	$\text{Fe}_7^{2+}\text{S}_8$	x	
Mackinawite	Sulfide	Mck	$(\text{Fe}, \text{Ni})_1 + x\text{S}$	x	
Pentlandite	Sulfide	Pn	$(\text{Fe}, \text{Ni})_9\text{S}_8$	x	
Cubanite	Sulfide	Cbn	CuFe_2S_3	x	
<i>Fe(II) compounds</i>					
Almandine	Silicate-garnet	Alm	$\text{Fe}_3^{2+}\text{Al}_2\text{Si}_3\text{O}_{12}$	x	
Augite	Silicate-clinopyroxene	Aug	$(\text{Ca}, \text{Na})(\text{Mg}, \text{Fe}^{2+}, \text{Al}, \text{Fe}^{3+}, \text{Ti})[(\text{Si}, \text{Al})_2\text{O}_6]$	x	x
Biotite	Silicate-phyllsilicate-mica	Bt	$\text{KMg}_{2.5}\text{Fe}_{0.5}\text{AlSi}_3\text{O}_{10}(\text{OH})_{1.75}\text{F}_{0.25}$	x	x
Chlorite (Ripidolite) (CCa2)	Silicate-phyllsilicate-chlorite	CCa2	$(\text{Mg}, \text{Fe}, \text{Al})_6(\text{Al}, \text{Si})_4\text{O}_{10}(\text{OH})_8$		x
Fayalite100 (Fo40)	Silicate-olivine	Fo0	$\text{Fe}_2^{2+}\text{SiO}_4$	x	
Fayalite60 (Fo40)	Silicate-olivine	Fo40	$\text{Fe}_2^{2+}\text{SiO}_4\text{--Mg}_2\text{SiO}_4$	x	
Fayalite20 (Fo40)	Silicate-olivine	Fo60	$\text{Fe}_2^{2+}\text{SiO}_4\text{--Mg}_2\text{SiO}_4$	x	
Ferrosilite	Silicate-orthopyroxene	Fs	$\text{Fe}^{2+}\text{MgSi}_2\text{O}_6$	x	
Green Rust-Cl, synthetic	Oxide-oxyhydroxide	GRCI	$\text{Fe}_3^{2+}\text{Fe}^{3+}(\text{OH})_8\text{Cl}\cdot\text{H}_2\text{O}$	x	
Green Rust-CO ₃ , synthetic	Oxide-oxyhydroxide	GRCO3	$\text{Fe}_3^{2+}\text{Fe}^{3+}(\text{OH})_{12}\text{CO}_3\cdot 2\text{H}_2\text{O}$	x	x
Green Rust-SO ₄ , synthetic	Oxide-oxyhydroxide	GRSO4	$\text{Fe}_3^{2+}\text{Fe}^{3+}(\text{OH})_{12}\text{SO}_4\cdot 2\text{H}_2\text{O}$	x	x
Mauna Loa Glass	Silicate-glass	Gla	N/A	x	
Hercynite	Oxide-Spinel	Hc	$\text{Fe}^{2+}\text{Al}_2\text{O}_4$	x	
Hypersthene	Silicate-orthopyroxene	Hyp	$\text{Mg}^{2+}, \text{Fe}^{2+}\text{SiO}_3$	x	
Ilmenite	Oxide-Ti oxide	Ilm	$\text{Fe}^{2+}\text{TiO}_3$	x	
Iron sulfate	Sulfate salt	SO4	$\text{Fe}^{2+}\text{SO}_4$	x	x
Magnetite	Oxide-Spinel	Mag	$\text{Fe}_2^{+3}\text{Fe}^{2+}\text{O}_4$	x	x
Pigeonite	Silicate-clinopyroxene	Pgt	$\text{Mg}_{1.35}\text{Fe}_{0.55}^{2+}\text{Ca}_{0.1}\text{Si}_2\text{O}_6$	x	
Siderite	Carbonate	Sd	$\text{Fe}^{2+}\text{CO}_3$	x	
Vivianite	Phosphate	Viv	$\text{Fe}_3^{+2}(\text{PO}_4)_2\cdot 8\text{H}_2\text{O}$	x	
<i>Fe(III) compounds</i>					
Aegirine powder avg	Silicate-clinopyroxene	Aeg	$\text{NaFe}^{3+}(\text{Si}_2\text{O}_6)$	x	
Akaganeite	Oxide-oxyhydroxide	Aka	$\beta\text{-Fe}^{3+}\text{O}(\text{OH}, \text{Cl})$	x	x
Andradite	Silicate-garnet	Adr	$\text{Ca}_3\text{Fe}_2^{3+}\text{Si}_3\text{O}_{12}$	x	
Biogenic oxide	Oxide-oxyhydroxide	Box	N/A	x	x
Ca montmorillonite (STx-1)	Silicate-phyllsilicate-clay	STx	N/A	x	
Fe Alginate	Organic complex	Alg	$\text{Fe}-(\text{C}_6\text{H}_8\text{O}_6)_n$	x	
Fe Dextran	Organic complex	Dex	$\text{Fe-H}(\text{C}_6\text{H}_{10}\text{O}_5)_x\text{OH}$	x	
Feroxyhyte	Oxide-oxyhydroxide	Fxy	$\delta\text{-Fe}^{3+}\text{O}(\text{OH})$		
Ferrihydrite (2-line)	Oxide-oxyhydroxide	Fh2L	$\text{Fe}_3^{3+}\text{HO}_8\cdot 4\text{H}_2\text{O}$ approx	x	x
Ferrihydrite (6-line)	Oxide-oxyhydroxide	Fh6L	$\text{Fe}_3^{3+}\text{HO}_8\cdot 4\text{H}_2\text{O}$ approx	x	x
Ferrosmeectite	Aluminosilicate clay	Sme	N/A	x	x
Goethite	Oxide-hydroxide	Gt	$\alpha\text{-Fe}^{3+}\text{O}(\text{OH})$	x	x
Hematite	Oxide	Ht	$\alpha\text{-Fe}_2^{3+}\text{O}_3$	x	

Table 4
Error estimates for Fe species (see Section 3.5).

	Description of error	How determined	RSD _{Fe(II)} (%)	RSD _{Fe(III)} (%)	RSD _{Fe-sulfide} (%)	RSD _{FeTotal} (%)
<i>RSD_{dist}</i>	Variability in distribution of Fe on filter	GT2276 rerun during 2014–04	9	3	0.1	4
<i>RSD_{2014beam}</i>	Variability between beamline visits in 2014	GT2326 run 2014–04, 2014–09	16	7	8	11
<i>RSD_{2014-15beam}</i>	Variability between beamline visits in 2014–15	GT2443 run 2014–04, 2014–09	53	38	63	30
<i>RSD_{2015corr1}</i>	Variability between beamline visits in 2014–15, after correction 1	GT2443 run 2014–09, 2015–10	16	9	27	19

total Fe, respectively (Table 4). When we applied this approach to a sample measured in the 2014–09 and 2015–10 beamline visits (GT2443—see Table 3), however, we found a much greater difference (*RSD_{2014-15beam}* of 38–63% – Table 4). All samples run in 2015 had anomalously high proportions of reduced Fe compared to samples run in 2014. In 2015, there were problems with a hydrocarbon residue coating on the vertical collimator. The flare coming off this mirror increased the divergence onto the monochromator, yielding a degraded energy resolution that resulted in a vertical beam flare, which is erroneously detected as X-ray absorption at lower energies and thus interpreted as the presence of reduced Fe. This is clearly visualized by plotting the proportion of Fe(III), Fe(II), and Fe-sulfide species for Station 5, which were mostly collected in 2015, compared to Stations 2 and 3, which were collected in 2014 (Fig. 2).

A beam flare issue would also appear in XANES spectra as a shift to lower energies, so using reference XANES spectra collected under the same beamline conditions in the

chemical map calculations should account for the artifact. To correct for the flare-induced erroneous attribution of Fe to reduced species in 2015, we used a XANES spectrum taken from a sample in 2015 that appeared the most oxidized as an approximation of what a pure Fe(III) reference XANES would look like with the beam flare. The chemical map calculation was made with this oxidized sample XANES spectrum instead of HFO as an Fe(III) reference.

This approach yielded Fe species profiles that were shifted towards more oxidized Fe throughout the water column compared to the raw profiles (Fig. 2), and closer to the repeat sample (GT2443 from Station 5, 200 m) that was also run in 2014. Using the repeat sample, the relative standard deviations between the 2014 and corrected 2015 data, *RSD_{2015corr1}* (16% Fe(II), 9% Fe(III), 27% Fe sulfide) were much more similar to the typical variability encountered between beamline visits in 2014 (*RSD_{2014beam}*) (Table 4).

4. RESULTS

4.1. Oxygen and nitrite distribution

The stations reported here are the 11 stations sampled along 12°S closest to the Peruvian shelf (Fig. 1), which all showed a distinct ODZ (Figs. 3 and 4). At Stations 2 and 3 on the shelf, oxygen decreased from the surface until the detection limit was reached around 70 m and 35 m, respectively (Figs. 3 and 4). At these 2 stations the ODZ reached the ocean floor (122 m and 180 m). The upper boundary of the ODZ was around 65 m at Station 4, deepening westward to 200 m by Station 11, while the lower bound of the ODZ shoaled westward from about 500 m at Station 4 to about 240 m by Station 11. Oxygen levels were similar to those previously reported for the region by Hong and Kester (1986), Noffke et al. (2012), Vedamati et al. (2014), Chever et al. (2015). The standard methods to determine oxygen concentrations used here have detection limits of 1–2 μM and the nominal concentrations of oxygen within the ODZ were at or below this value. However, for reasons outlined previously (cf. Ulloa et al., 2012; Tiano et al., 2014), we assume oxygen concentrations are low enough that it is not important as an oxidant, if not completely absent.

Nitrite concentrations were highest near the shelf, reaching $\sim 10 \mu\text{M}$ at Stations 3, 4, and 5 (Figs. 3 and 4), and typically peaked in the upper part of the ODZ, with the depth

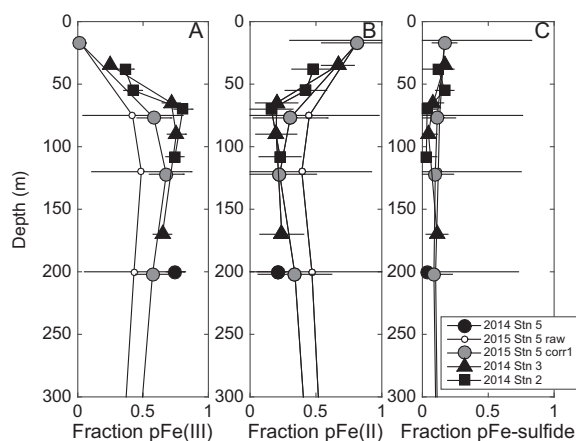


Fig. 2. Fractional concentrations of pFe species for three margin stations. (A) Fraction Fe(III), (B) Fraction Fe(II), (C) Fraction Fe sulfide for Stations 2 (squares), 3 (triangles) and 5 (circles). All data collected in 2014 are filled black symbols. Raw data collected in 2015 are small open symbols, and corrected data (see Section 3.5) are filled grey symbols. Error bars are standard deviations calculated from *RSD_{beam}* for Stations 2, 3, *RSD_{2014-15beam}* for Station 5 raw, and from *RSD_{2015corr1}* for Station 5 corrected (Tables 3 and 4).

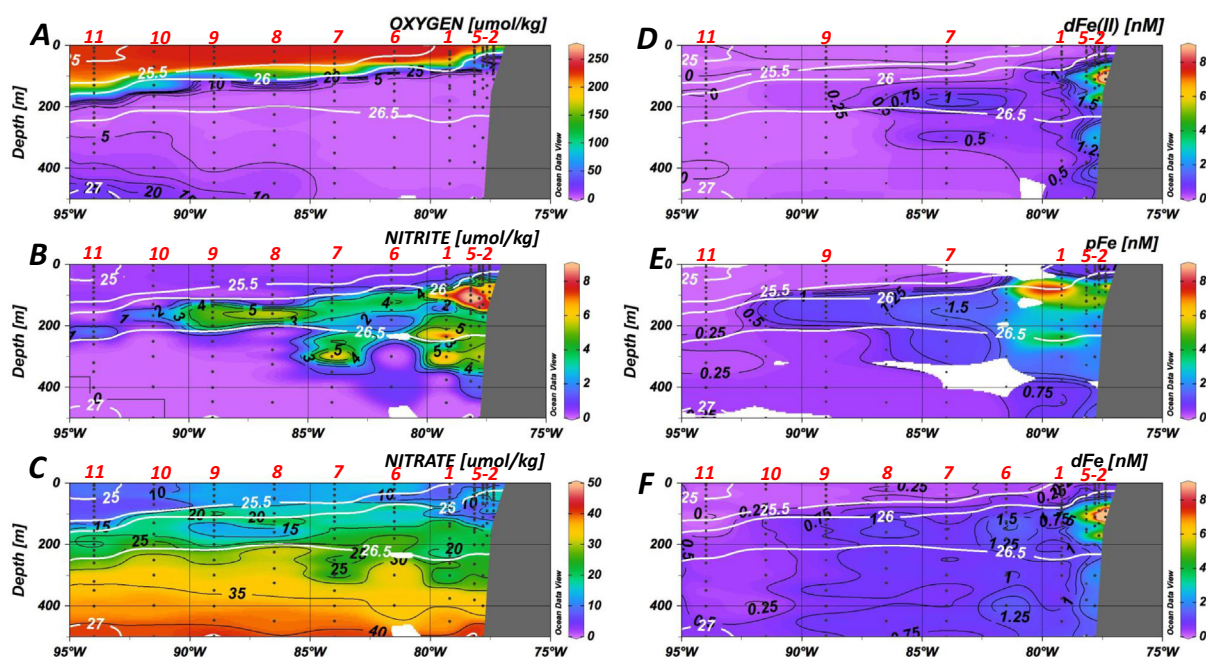


Fig. 3. Upper 500 m sections of OMZ study area indicated in Fig. 1. Sections show concentrations of (A) oxygen ($5 \mu\text{M}$ black contours up to $25 \mu\text{M}$), (B) nitrite ($1 \mu\text{M}$ black contours to $5 \mu\text{M}$), (C) nitrate ($5 \mu\text{M}$ black contours), (D) dissolved Fe(II) (0.25 nM black contours to 1.5 nM), (E) total particulate Fe (0.25 nM black contours to 1.5 nM), and (F) dissolved Fe (0.25 nM black contours to 1.5 nM). Potential density ($\sigma_{\theta,0}$) overlaid on each panel (0.5 kg/m^3 white contours). Above each section plot, station numbers in red indicate where data were available.

of the peak deepening westward from 69 m at Station 2 closest to the coast to 220 m at Station 11 furthest offshore, following the upper boundary of the ODZ (Figs. 3 and 4). At some stations (Stations 3, 5, 1, 7), there was more than one nitrite maximum within the ODZ (Fig. 4). As the top of the ODZ deepened westward, a distinct primary nitrite peak above the ODZ was sometimes also noticeable (e.g., $0.5 \mu\text{M}$ at 110 m at Station 11). Nitrite was undetectable below the ODZ (Figs. 3 and 4).

4.2. Distributions of dissolved Fe, dissolved Fe(II) and total pFe concentrations

DFe(II) concentrations were highest on the shelf, decreasing offshore (Fig. 3D). Vertical profiles of dFe(II) (Fig. 4) were generally characterized by maxima between 100 and 400 m within the ODZ, with the highest concentrations found in near-bottom samples on the shelf-slope including 25.6 nM at 108 m (bottom = 122 m) at Station 2 and 8.1 nM at 170 m (bottom = 180 m) at Station 3. Elevated ($\geq 1 \text{ nM}$) concentrations of dFe(II) were generally confined to Stations 2–5 on the shelf/slope. DFe(II) remained detectable within the ODZ west of Station 5, primarily within the secondary nitrite maximum until that feature disappeared west of Station 11.

Concentrations of total dissolved Fe (dFe) were also highest on the shelf and decreased offshore, with a sharp drop west (offshore) of Station 5, as for dFe(II). West of Station 5, however, the vertical distribution of dFe (Fig. 3F) did not exhibit a maximum in the ODZ, but

rather increased with depth below the ODZ. Within the upper part of the ODZ, dFe(II) generally accounted for a majority of the total dissolved Fe, but was a very small percentage in the lower part of the ODZ and below the ODZ.

Concentrations of total pFe were generally highest near the shelf and slope stations east of Station 1 (Figs. 3 and 4). Vertical profiles of pFe in the upper 500 m were characterized by a near-surface minimum, and a local maximum in the upper part of the ODZ. The near-surface ($<50 \text{ m}$) minima ranged from $1.2 \pm 0.09 \text{ nM}$ at 38 m at Station 2 closest to the coast to below the detection limit ($<0.14 \text{ nM}$) west of Station 7. The pFe maxima in the upper ODZ deepened westward, and ranged from a maximum of $10.9 \pm 0.59 \text{ nM}$ at 70 m at Station 2 closest to the coast to $0.4 \pm 0.05 \text{ nM}$ at 225 m at Station 11 offshore. The maximum concentrations of pFe in the ODZ remained relatively high ($>3 \text{ nM}$) from Station 2 west through Station 1, before decreasing more quickly westward starting at Station 7.

4.3. Synchrotron analysis of pFe speciation

4.3.1. Redox state of pFe in the ODZ

Synchrotron-based chemical species maps of 1–51 μm particles from the upper 600 m of three coastal stations (Stations 2, 3, 5) show that the samples in the pFe maxima were dominated by Fe(III) (blue color in Fig. 5). Quantification of the intensities in each Fe species channel show that the percentage of total pFe as Fe(III), Fe(II), and Fe-sulfide exhibited similar profiles at the three stations. All three stations were characterized by a minimum in %

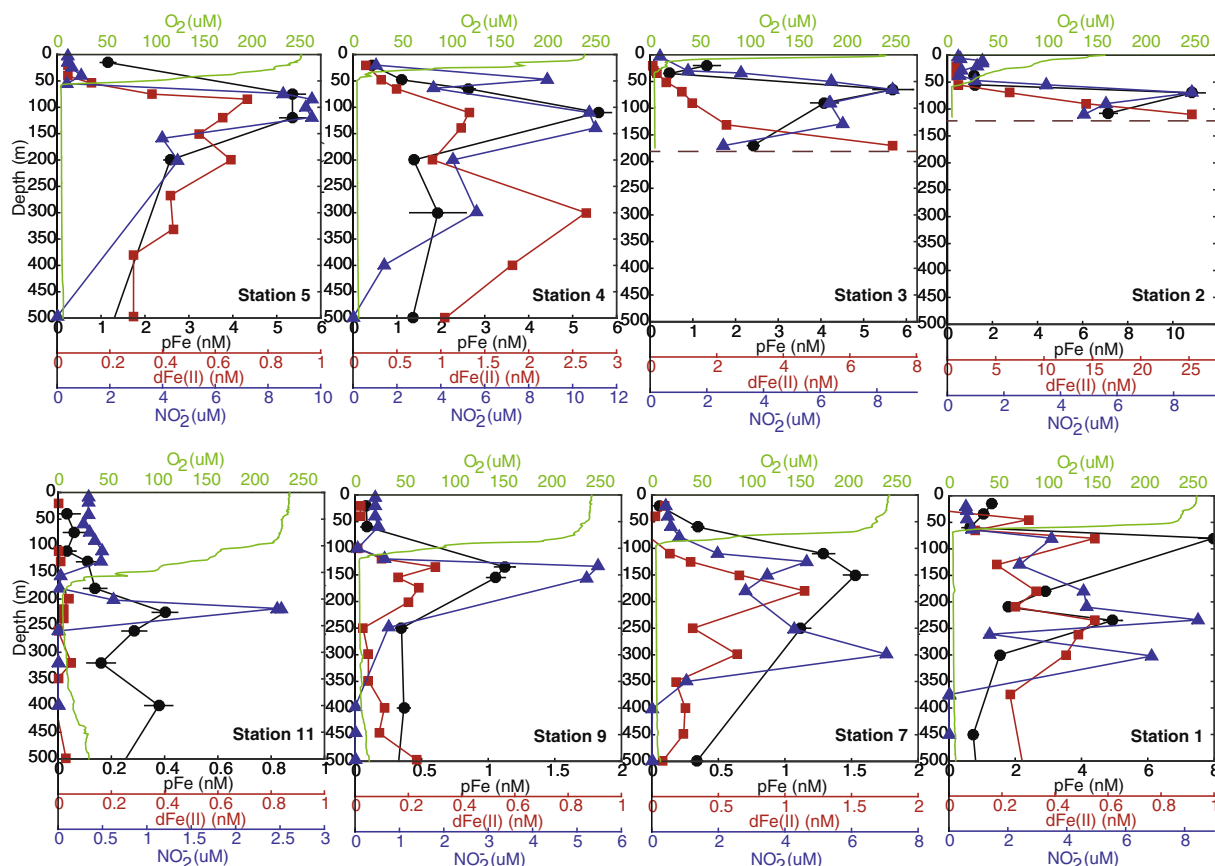


Fig. 4. Profiles of total pFe by ICP-MS (black circles), dFe(II) (red squares), nitrite (blue triangles) and oxygen (green line) for the upper 500 m. Margin stations on top row, and open ocean stations on bottom row, arranged from east on right to west on left. Horizontal dashed line in Station 2 and 3 panels indicates water depth. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

Fe(III) near surface at each of the three stations where total pFe was very low (Figs. 2 and 6), increasing with depth until a maximum of up to 80% Fe(III) was reached between 50 m and 150 m (Fig. 2).

Total pFe determined by XRF was generally comparable to total pFe determined by ICP-MS (Fig. 6). The sample areas from which XRF pFe were determined represented only about 8 mL of equivalent volume filtered, compared to 25–50 L for ICP-MS subsamples, so the ICP-MS-derived pFe are expected to be more representative.

All pFe species were low in near surface waters, and increased in concentration from the surface into the upper part of the ODZ (Fig. 6). In the euphotic zone, the particulate Fe:P ratio approached typical stoichiometric ratios found in phytoplankton (Fig. S1), suggesting that a significant fraction of pFe there may have been biologically-bound. Below the euphotic zone, particulate Fe:P ratios greatly exceed phytoplankton ratios.

Overall, Fe-sulfides were a minor component of pFe in the water column (Fig. 6), and were the greatest fraction of total pFe (up to 17%) above 60 m in the euphotic zone (Fig. 2). Particulate acid volatile sulfides (AVS) were measured on separate pump subsamples and only detected in the euphotic zone (Ohnemus et al., 2016), suggesting a phy-

toplankton source for the AVS and perhaps the Fe-sulfides detected here in the euphotic zone. Pyrite particles are identifiable in chemical maps as intense, small red spots, and were occasionally detected below the euphotic zone at Stations 2 (70 m), 3 (170 m), and 5 (120 m) (Fig. 5) and confirmed by their distinctive XANES spectra (Fig. 7). The abundance of distinct pyrite particles was very low, however, and they may have derived from resuspended and laterally-transported pyrite formed in sediments (*cf.* Lam et al., 2012).

The majority of euphotic zone pFe was categorized as Fe(II), and may represent biologically-bound Fe, given the Fe:P ratios at these depths. Below the euphotic zone, most of the total pFe, including the increase in the upper ODZ, was comprised of pFe(III). Particulate Fe:Al exceed crustal ratios throughout the upper 500 m (Fig. S2). Particulate Fe below the euphotic zone is therefore unlikely to be dominated by either biologically bound or silicate-bound components.

4.3.2. Speciation of pFe

We performed X-ray Absorption Near-Edge Structure (XANES) spectroscopy on multiple Fe-rich spots from each chemical species map to confirm the categorization of pFe

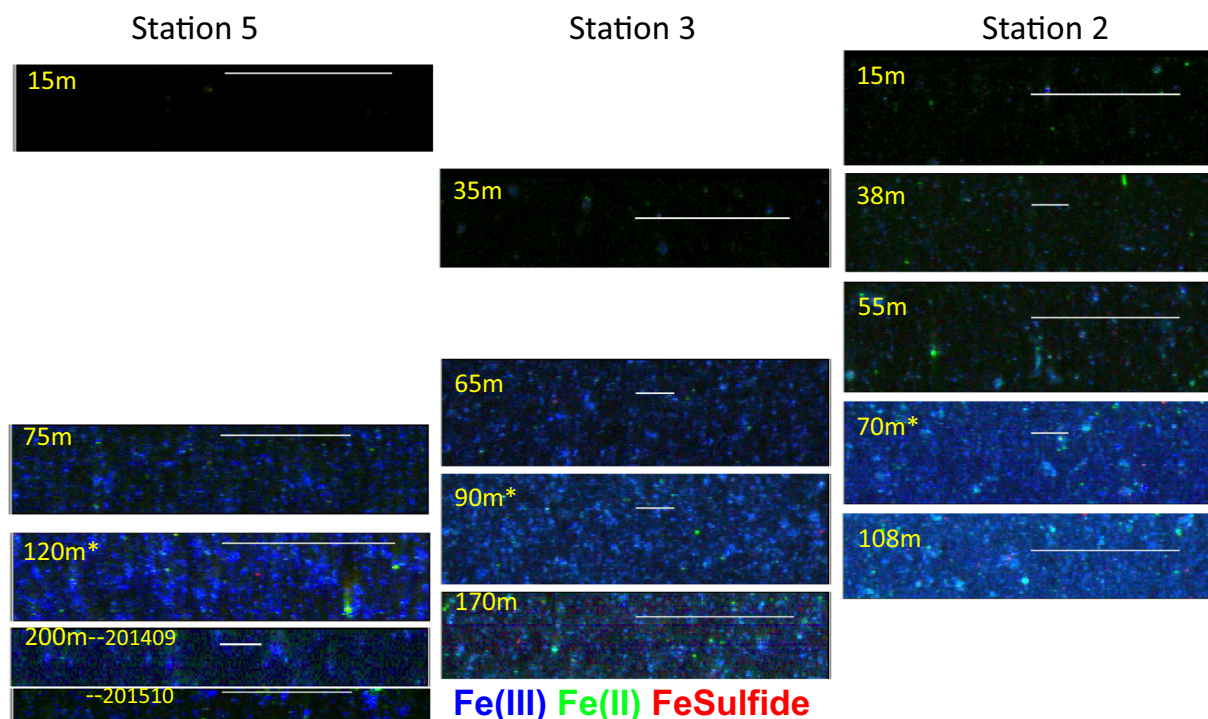


Fig. 5. Tricolor synchrotron chemical maps showing the distribution of pFe species; blue color shows Fe(III), green shows Fe(II) and red shows Fe Sulfide species. The white levels are set to the same values for the blue, green, and red channels, and are adjusted to the pixel density of the Fe standard (Table 1) to account for different detector configurations between beamline visits. The intensities of the colors in the blue, green, and red channels indicate the concentrations of Fe(III), Fe(II), and Fe sulfide species in each pixel. All maps are approximately 1000 μm across. White scale bars are automatically generated by the plotting program and are 70 μm , 100 μm , 299 μm or 400 μm (exact lengths indicated in Table 2). Depths of the ODZ pFe maxima are indicated by *. The predominant blue color of the chemical species maps at the pFe maxima indicate that pFe is present as mostly Fe(III).

redox state by the chemical mapping procedure, and also to further determine the chemical speciation of the pFe at a higher level of detail than is possible by chemical species mapping. Fe K-edge $\mu\text{-XANES}$ spectra of multiple spots from samples from the depth of the ODZ pFe maximum (Station 2–70 m, $n = 9$; Station 3–90 m, $n = 9$; Station 5–120 m, $n = 15$) were compared to the XANES spectra of model Fe species (HFO-Fe4-P4 for Fe(III), FeSO_4 for Fe(II), and pyrite for Fe-sulfide) and used to calculate the chemical species maps (Fig. 7). With the exception of two spectra from Station 5 that appear significantly chemically reduced, the 31 other XANES spectra plotted close to the Fe(III) model spectrum, confirming the dominance of Fe(III) predicted by the chemical species map at the ODZ pFe maximum. Indeed, a composite model Fe XANES spectrum that was a 25%/75% mixture of the Fe(II) and Fe(III) model species, which was the overall oxidation state predicted by the chemical species maps at the ODZ pFe maximum (Fig. 2), is an excellent approximation of the ensemble of sample XANES spectra (Fig. 7). The two spectra that appear chemically reduced were identified in the chemical species map as Fe particles having significant pyrite and Fe(II), consistent with their XANES spectra (Fig. 7).

Because the detailed identification of poorly crystalline Fe(III) species is difficult by XANES, we chose two Fe

(III)-rich spots on samples within the ODZ to conduct EXAFS analysis, one from Station 2 at 108 m, which was near bottom (bottom depth = 122 m), and one from Station 5 at 200 m (bottom depth = 2072 m) (Fig. 8). Both samples were well within the ODZ, and many tens of meters away from the oxycline, making it unlikely that these samples were exposed to oxygen in situ. The EXAFS of the Fe(III)-rich spot from Station 5 at 200 m was well fit with two Fe(III) oxyhydroxides: 34% lepidocrocite ($\gamma\text{-FeOOH}$) and 74% feroxyhyte ($\delta'\text{-FeOOH}$). The EXAFS of the Fe(III)-rich spot from Station 2 at 108 m (near-bottom) was well fit with two Fe(III) oxyhydroxides (20% lepidocrocite and 25% poorly crystalline hydrous ferric oxide (HFO), and 46% Fe(III)-containing clay (illite).

5. DISCUSSION

5.1. Sources of dFe(II) in the ODZ

While there have been episodic observations of sulfide accumulation in the Eastern Tropical Pacific ODZ, especially on the shelf (Dugdale et al., 1977; Bruchert et al., 2003), and a “cryptic sulfur cycle” in which sulfide production and sulfate reduction are tightly coupled has been proposed (Canfield et al., 2010), the redox potential of the Eastern Tropical Pacific is generally not low enough to be

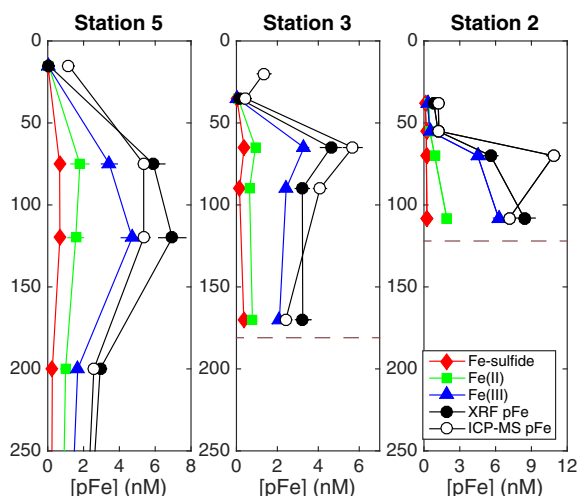


Fig. 6. Absolute concentrations of total pFe determined by XRF (black circles) or ICP-MS (white circles), and pFe species from chemical species mapping: Fe(III) (blue triangles), Fe(II) (green squares), Fe-sulfide (red diamonds) for Stations 5, 3, and 2. Error bars for XRF Fe and Fe species are one standard deviation, calculated from RSD_{beam} (Table 4) for Stations 2, 3 collected in 2014, and from RSD_{corr} (Table 4) for Station 5 collected mostly in 2015. Error bars for ICP-MS Fe are propagated from uncertainty in the blanks and repeat analyses. Horizontal dashed line in Station 2 and 3 panels indicates water depth. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

sulfidic. Measurement of multiple redox-sensitive trace elements (iodine (I), manganese (Mn), chromium (Cr), selenium (Se), and iron (Fe)) in oxygen deficient waters of the Eastern Tropical North Pacific suggested the redox potential was set by the nitrate-nitrite, Cr(VI)—Cr(III), and selenate-selenite redox couples, well above that required for in-situ dissimilatory Fe or sulfate reduction (Rue et al., 1997). Similar measurements during the EPZT cruise also indicate that the bulk water column redox potential was not low enough for in-situ Fe(III) reduction (Cutter and Nielsdottir, 2014). Observed dFe(II) in the ODZ water column may have originated from sediments, where the redox potential is low enough for benthic Fe reduction, and transported along isopycnal surfaces. Indeed, the $26.1\text{--}26.55\text{ kg/m}^3$ potential density surface connects shelf sediments to the elevated dFe(II) features in the upper ODZ (cf. Fig. 3). A sedimentary source of dFe(II) is supported by two lines of evidence. First, light Fe isotopic values, a tracer of dissimilatory Fe(III) reduction, were observed in the leachable particulate Fe pools in the ODZ zone of the EPZT cruise (Marsay and John, 2014), as well as in shelf waters in a previous study along the Peruvian upwelling coast (Chever et al., 2015). Second, measurements of Ra-228, a tracer of recent (<10 years) sedimentary contact, revealed surface plumes of water with a strong shelf input that decline sharply above the oxycline, but are still elevated above background at the depths of the dFe(II) features (Sanial et al., 2017). Laterally-transported

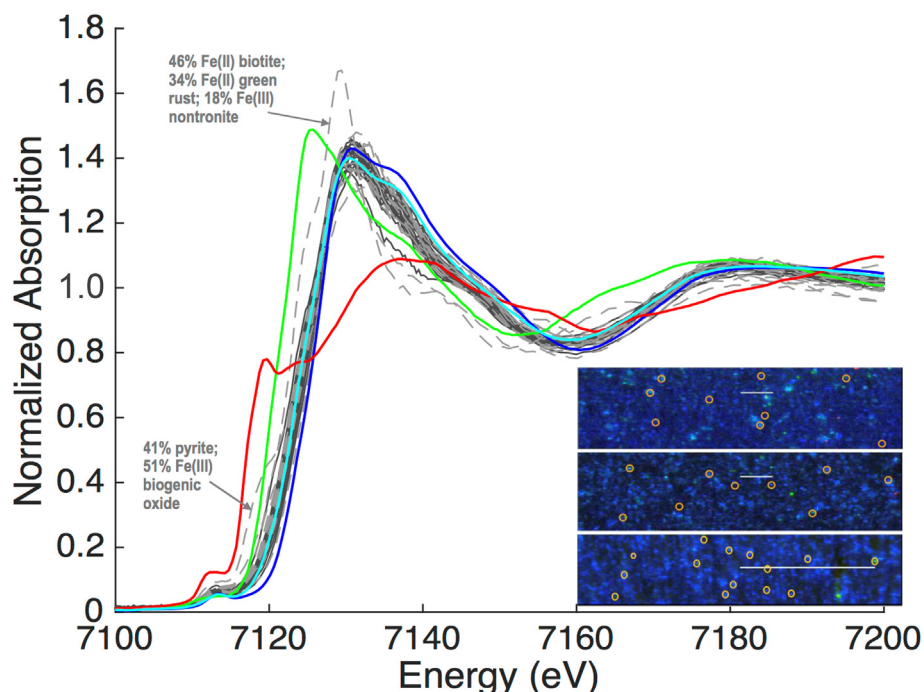


Fig. 7. Fe K-edge μ -XANES spectra of multiple spots from samples from the depth of the ODZ pFe max (grey), overlain by model Fe species (colors). Locations of μ -XANES indicated as orange circles in chemical species maps in inset. Black, solid lines are 18 spectra collected in 2014 (Stn 2–70 m, $n = 9$, top map and Stn3–90 m, $n = 9$, middle map); light grey, dashed lines are 15 spectra collected in 2015 (Station 5–120 m, bottom map); red, green, and dark blue lines are the Fe sulfide (pyrite), Fe(II) (FeSO_4), and Fe(III) (HFO) model Fe species, respectively, used in the chemical mapping calculation; the light blue line is a 25%/75% mixture of the Fe(II) and Fe(III) model species which is a good approximation of most of the XANES spectra from the pFe maximum, and consistent with the quantification of Fe redox state by chemical mapping (Fig. 5). The fits for the only two spectra that are significantly chemically reduced are indicated.

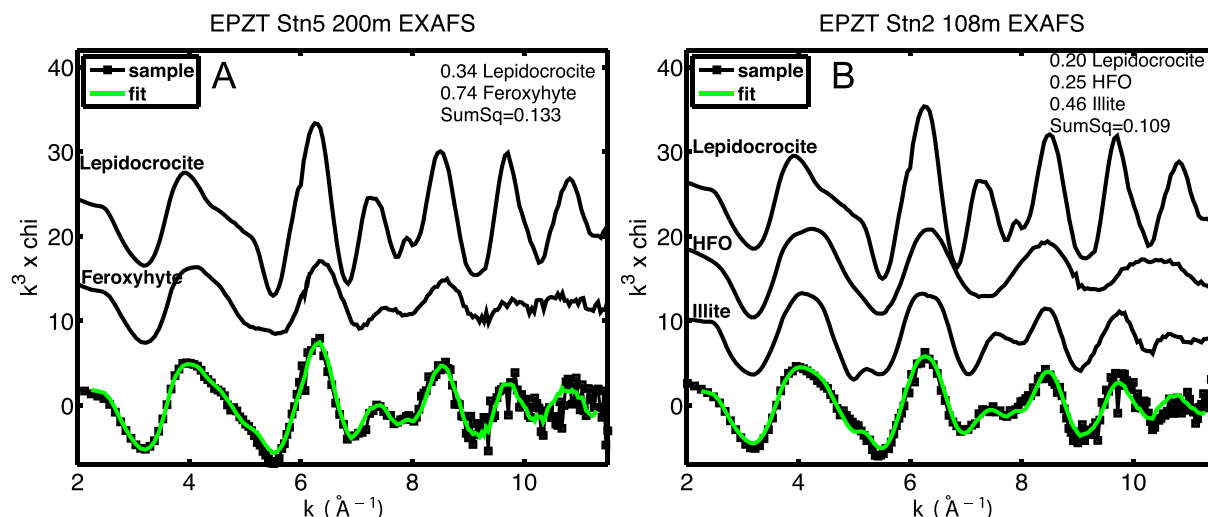


Fig. 8. Fe K-edge μ -EXAFS spectroscopy of an iron-rich Fe(III) spot from Station 5, 200 m (A), and from Station 2, 108 m (B), both within the oxygen deficient zone. The best fit (green line) to the sample spectra (open squares) is indicated. EXAFS spectra of model Fe species appearing in the fit also indicated (solid lines above the sample spectra and fits) (See Table 2 for description of model species). (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

shelf sources of dFe(II) may also be supplemented by in-situ sources of dFe(II). Although the bulk redox potential is not low enough for in-situ Fe(III) reduction, dissimilatory in-situ Fe(III) reduction could potentially occur in microenvironments (*cf.* Canfield et al., 2010). Moffett et al. (2007) also hypothesized that in situ reduction of Fe (III) to Fe(II) might occur as part of an assimilatory Fe acquisition strategy. Finally, as we will further discuss in Section 5.3, some dFe(II) could also be produced in-situ from the remineralization of particulate organic matter.

5.2. Conversion of dFe(II) to pFe in the ODZ of the EPZT

In the upper ODZ of the EPZT (defined here as the 26.1–26.55 kg/m^3 potential density surfaces), depth-

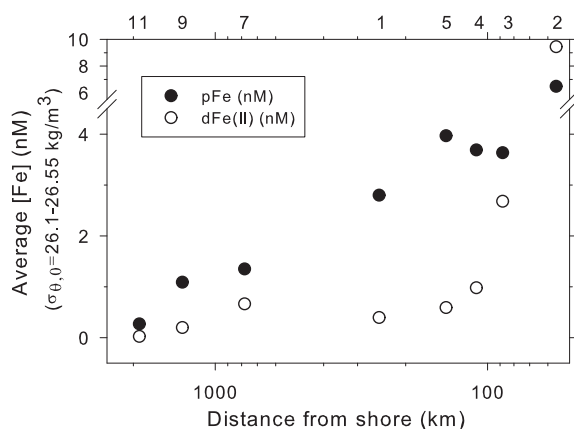


Fig. 9. Average concentrations of pFe (closed circles) and dFe(II) (open circles) as a function of distance from shore, plotted on a log₁₀ scale. Average concentrations are depth weighted for samples in the upper ODZ (defined here as $\sigma_{\theta,0}$ between the 26.1–26.55 kg/m^3). Station numbers indicated at top of figure.

averaged dissolved Fe(II) concentrations drop off more quickly with distance from shore than pFe, especially between Stations 2–5 (Fig. 9), consistent with a conversion of dFe(II) to pFe. This confirms earlier hypotheses that dFe (II) is lost to the particulate phase within the Peruvian ODZ (Vedamati et al., 2014). Prior to the cruise, we anticipated that particulate Fe would be dominated by weakly-adsorbed Fe(II) in the OMZ and by Fe(III) oxyhydroxides at the oxygen gradients (horizontal and vertical), but our synchrotron-based measurements of the speciation of pFe suggest otherwise.

We find that pFe is overwhelmingly dominated by the Fe(III) redox state below the euphotic zone, even in the ODZ (Figs. 2, 5 and 6). This was confirmed by chemical species mapping (Fig. 5), and XANES analyses of dozens of individual Fe(III) spots (Fig. 7). The detailed EXAFS analyses of two individual Fe(III) spots confirmed that the Fe(III) identified by chemical species mapping and XANES was present as Fe(III) oxyhydroxides in the case of the mid-water column sample (Station 5, 200 m), and as a mixture of Fe(III) oxyhydroxides and an Fe(III)-containing clay mineral, consistent with the clay mineral illite, in the case of the near-bottom sample (Station 2, 108 m) (Fig. 8). As the near-bottom sample was within 15 m of the bottom, the illite was likely from sediment resuspension.

The identification of the specific Fe(III) species by EXAFS was critical for assessing whether the presence of pFe(III) resulted from in-situ oxidation in the ODZ, or was simply a result of unavoidable oxidization upon sample recovery. Importantly, both EXAFS required small but significant proportions of the mineral lepidocrocite (γ -Fe(III) $\text{O}(\text{OH})$) in the fits. The EXAFS of lepidocrocite shows high amplitude around $k = 6.2 \text{ \AA}^{-1}$, which was a distinctive feature in both EXAFS sample spectra (Fig. 8). Indeed, the ten best linear-combination fits all required lepidocrocite. Lepi-

lepidocrocite is a crystalline Fe oxyhydroxide that is typically formed underwater by oxidation of dissolved Fe(II) via a green rust (mixed Fe(II)-Fe(III) hydroxide) intermediate (Schwertmann and Fechter, 1994). Because lepidocrocite forms by oxidation of aqueous Fe(II) solutions (Cornell and Schwertmann, 2003), and samples used for pFe speciation were collected by in-situ filtration, in which particles are already separated from the aqueous phase at the time of recovery and exposure to air, the presence of lepidocrocite in the ODZ samples is strong evidence that it was formed in-situ in the ODZ, and is not a preservation artifact.

The presence of hydrous ferric oxides (HFO) in the Station 2, 108 m EXAFS can also be consistent with in-situ production. Even though poorly crystalline HFO and ferrihydrite typically form from the rapid hydrolysis of aqueous Fe(III) solutions (Cornell and Schwertmann, 2003), these mineral forms can also result from the interference of other solutes in the formation of lepidocrocite from aqueous Fe(II) solutions. For example, the presence of silicate, phosphate, and calcium ions has been shown to hinder the formation of lepidocrocite from oxidation of dFe(II), resulting instead in the formation of poorly crystalline hydrous ferric oxides (HFO) or ferrihydrite (Schwertmann and Thalmann, 1976; Voegelin et al., 2010, 2013; Senn et al., 2015). As these solutes are orders of magnitude higher in concentration than Fe in Peruvian ODZ waters, it is reasonable to expect that HFO could have formed from the oxidation of dFe(II) instead of or in addition to lepidocrocite. HFO was required in the top three linear-combination fits in the Station 2 sample, and in the second and third best two-component linear-combination fits in the Station 5 sample.

Finally, we cannot rule out the possibility that some of the pFe(III) species observed in our samples are a result of oxidation upon sample recovery. Feroxyhyte (δ -Fe(III) O(OH)) was required in the best 2-component fit for the Station 5 sample (Fig. 8a), and was required in nine of the top ten 3-component fits for this sample (not shown). While feroxyhyte is found in deep sea iron-manganese nodules where it forms under high pressure conditions (Chukhrov et al., 1977), it can also be formed by extremely rapid oxidation of an Fe(II)-containing hydroxide such as green rust upon exposure to air (Carlson and Schwertmann, 1980; Cornell and Schwertmann, 2003). The feroxyhyte detected in our sample may have therefore formed from the rapid oxidation of green rust upon exposure to air on sample recovery. So while feroxyhyte detected in our samples may not have formed in-situ, its precursor, green rust, is formed by direct precipitation of aqueous Fe(II) upon oxidation (Schwertmann and Fechter, 1994) and must have formed in-situ.

Overall, the robust presence of lepidocrocite in our data provides evidence that at least some of the Fe(III) oxyhydroxides in our samples were oxidized in-situ and not simply by exposure of the particulate samples to oxic conditions during sample recovery and handling, especially since samples were protected from oxidation during shipping, storage, and XAS analysis. The other pFe(III) species observed also support the idea that dFe(II) is lost by partial (to green rust) or complete (to lepidocrocite or HFO)

oxidation in the ODZ, rather than only by adsorption of dFe(II) to particle surfaces. In contrast to our original expectation, we saw no indication of a rise in pFe at the outer edges of the OMZ as the oxygen levels started to increase, and in fact observed the highest pFe concentrations in the upper part of the ODZ (Fig. 3).

5.3. Linkages between the Fe and N cycles in OMZs

We have argued that oxygen is absent in the ODZ. The other important oxidant in the euphotic zone, hydrogen peroxide (Moffett and Zika, 1987), is also absent below the mixed layer in the Peru upwelling region (Zika et al., 1985). Therefore, nitrate or nitrite is the most likely oxidant. The vertical and horizontal distributions of nitrite and total pFe show a striking similarity within the ODZ of the 2013 GEOTRACES EPZT cruise (Figs. 3 and 4). Indeed, pFe peaked at the same depth as nitrite at every station. A property-property plot of pFe and nitrite for samples between $\sigma_{\theta,0} = 26.1$ – 26.55 kg/m^3 , the upper part of the ODZ, for Stations 1–11 (Fig. 10) shows a significant correlation ($R^2 = 0.45$; $p = 0.0002$), with a best-fit slope of 0.60 nM pFe/ μM nitrite.

The correlation between pFe and nitrite suggests a role for nitrite as an oxidant for dFe(II). The abiotic oxidation of dFe(II) by nitrate (NO_3^-) or nitrite (NO_2^-) are both thermodynamically favorable, but kinetically inhibited and require a catalyst to proceed at the pH of seawater. These reactions have been studied in the laboratory since the 1960s: dFe(II) can be oxidized by NO_3^- to form Fe oxyhydroxide (α -FeOOH) and NO_2^- , N_2O , N_2 , or NH_4^+ or by NO_2^- to form N_2O , N_2 , or NH_4^+ (Buresh and Moraghan, 1976; Picardal, 2012 and references therein; Wulstein and Gilmour, 1966).

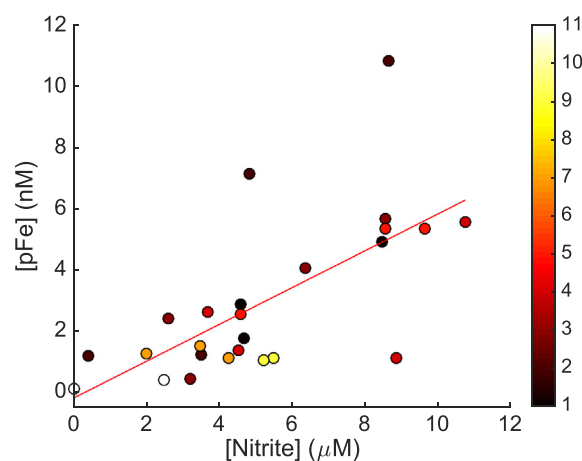


Fig. 10. Relationship between nitrite and particulate iron (pFe) for samples in the upper part of the ODZ between $\sigma_{\theta,0} = 26.1$ – 26.55 kg/m^3 , approximately 100–300 m, for Stations 1–11. The red line is the ordinary least squares regression fit: $R^2 = 0.45$; $p = 0.0002$. Colorbar indicates station number, with more margin-influenced stations in blue and more open ocean stations in red. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

Many catalysts for the abiotic oxidation of dFe(II) by nitrite have been identified, including Cu^{2+} (Moraghan and Buresh, 1977), Fe(II)-binding ligands such as nitrilotriacetic acid (NTA) (Kopf et al., 2013), or mineral surfaces such as Fe oxyhydroxides and clays (Van Cleemput and Baert, 1983; Sorensen and Thorling, 1991; Hansen et al., 1994; Coby and Picardal, 2005; Tai and Dempsey, 2009). Some studies of the oxidation of dFe(II) by nitrite have shown a strong autocatalytic effect, where the initial formation of Fe oxyhydroxides greatly accelerates subsequent oxidation kinetics (Jones et al., 2015; Buchwald et al., 2016). Given the dilute nature of the open ocean water column compared to the conditions under which these experiments were generally conducted (subnanomolar compared to millimolar), many of these mechanisms are unlikely to be important in the aqueous phase, but may become significant on particle surfaces.

In the late 1990s, microbially-mediated NO_3^- -dependent anaerobic Fe(II) oxidation (NDFO), an autotrophic metabolism that couples the reduction of nitrate to the oxidation of Fe, was discovered (Straub et al., 1996). Enrichment cultures have demonstrated that this appears to be a widespread metabolism and it has been identified in sediments from freshwater, brackish, and marine sediments (Benz et al., 1998). Recent studies have shown that it can be difficult to experimentally separate chemical oxidation of Fe(II) by microbially produced nitrite from microbial Fe(II) oxidation that is truly coupled to nitrate reduction (Picardal, 2012; Kopf et al., 2013). Indeed, most strains capable of anaerobic Fe(II) oxidation appear to be mixotrophic and require an organic carbon source to fully oxidize Fe(II) (Klueglein et al., 2015).

Recently, genomic data of near-bottom water from the Peruvian ODZ margin showed the presence of nitrate reducing bacteria that are related to a known Fe oxidizer (Scholz et al., 2016). The specific metabolism is not known, but it does point to the possibility of microbially-catalyzed Fe oxidation in ODZ waters. To our knowledge, the metabolism of putative anaerobic Fe oxidizers in the open ocean ODZ water column has not been investigated. While the details of the coupling of nitrate or nitrite reduction and Fe oxidation in the ODZ water column are unknown, the similar distributions of pFe and nitrite in our section (Figs. 4 and 10) strongly point to the importance of nitrate and nitrite for oxidizing Fe in the Peruvian ODZ.

Many Fe oxyhydroxide phases, including lepidocrocite, goethite, 2-line ferrihydrite, magnetite, maghemite, and green rust, have been identified as oxidation products of both abiotic (Jones et al., 2015) and biotic (Weber et al., 2006) anaerobic oxidation of Fe(II) conducted in controlled laboratory conditions. Other studies have shown that oxidation of dFe(II) by O_2 also produces a mixture of ferrihydrite, lepidocrocite, and goethite (Jones et al., 2014). Unfortunately, the identities of the mineral phases of the oxidation products alone do not seem to be sufficient to constrain whether the mechanism of oxidation was abiotic or microbial, nor do they constrain the nature of the oxidant.

Maxima in pFe and nitrite are also coincident with maxima in dFe(II) (Fig. 4). The coincidence of the secondary

nitrite maximum and dFe(II) has been previously reported in all three of the world's ODZs: Peru (Vedamati et al., 2014; Kondo and Moffett, 2015), Arabian Sea (Moffett et al., 2007; Kondo et al., 2010; Kondo and Moffett, 2013) and Mexican (Heller et al., 2013) ODZs. A similar feature was observed here albeit with less vertical resolution than in the previous studies.

One explanation for the coincidence of nitrite and dFe(II) is a vertical transport mechanism that couples the oxidation of organic matter by nitrate (nitrate reduction to nitrite) (Bange et al., 2005) with the regeneration of dFe(II) from organic matter (Kondo and Moffett, 2013). Euphotic zone pFe has significant biologically-bound Fe(II) (Fig. 2) (von der Heyden et al., 2014), indicating that there is a supply of reduced Fe to the ODZ from above. Regeneration of these particles in the secondary nitrite maximum could result in the accumulation of dFe(II) and nitrite. The secondary nitrite maximum is significant as a depth horizon characterized by high rates of microbial processes, including denitrification and anaerobic ammonium oxidation (anammox) (Dalsgaard et al., 2012). Accumulation of dFe(II) has been attributed to microbial processes associated with remineralization, dissimilatory reduction, or as part of an Fe acquisition strategy (Kondo and Moffett, 2015). The coincidence of dFe(II), nitrite and pFe(III) suggests that this is also a zone of dynamic redox cycling.

It is noteworthy that the oxidation state of pFe shifts from +2 in the euphotic zone to +3 in the upper ODZ. This suggests that at least some of the change could occur on the particles themselves. Consider a particle formed in the euphotic zone and remineralized in the secondary nitrite maximum. The valence of Fe initially is +2. Upon colonization by denitrifiers and through organic matter degradation, a fraction of this Fe is released as dFe(II) and diffuses away. However, a significant fraction may be oxidized along with the organic matter by denitrifiers (many of which can oxidize Fe(II) as well as organics) within the particle regime, thus remaining on the particles. However, there are reasons why this is likely to be of secondary importance, at least around the shelf/slope break. First, if the origin of the Fe was sinking particles, we would expect to also see significant Fe oxyhydroxide formation within the oxycline from the oxidation of regenerated Fe(II) by oxygen, yet we do not. Mesoscale eddies, which are widely reported in this area (cf. Chelton et al., 2011), could enhance lateral transport of dFe(II) fluxes from sediments. If there is lateral transport of a benthic dFe(II) source, then this dFe(II) source must be balanced by a sink, which could be the oxidation to the Fe oxyhydroxides we observed. The oxidation of a benthically derived dFe(II) is further supported by the light isotopic values of leachable pFe found in the ODZ (Marsay and John, 2014).

A schematic that summarizes the above couplings of the Fe and N redox cycles is shown in Fig. 11. DFe(II) in the ODZ can thus be supplied by three processes: isopycnal transport of dFe(II) from sediments, in situ dissimilatory iron reduction (DIR) in microenvironments of particles, or from remineralization of particulate organic matter. The oxidation of dFe(II) could be in the form of nitrate-

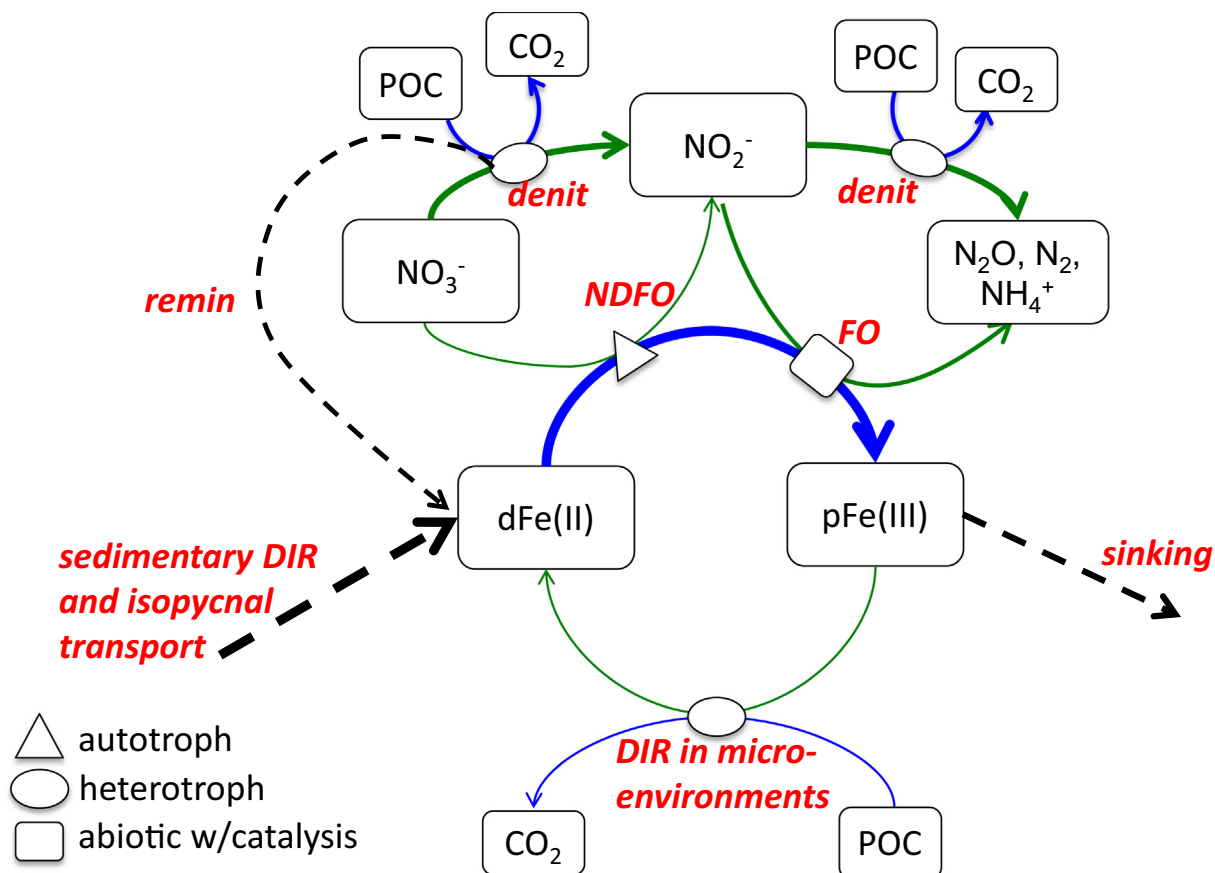


Fig. 11. Schematic illustrating the coupling of Fe and N redox cycling in the ODZ water column. Oxidation reactions are blue arrows; reduction reactions are green arrows. Redox reactions are mediated by microbial catalysts (triangles for autotroph/mixotroph; ovals for heterotrophs) or abiotic or unknown microbial catalysts (rectangle). Non redox processes that are sources or sinks of dFe(II) and pFe(III) are black dashed arrows. Thinner arrows represent more speculative reactions. Abbreviations are: NDFO—NO₃⁻-dependent Fe(II) oxidation; denit—denitrification; remin—remineralization; FO—Fe(II) oxidation; DIR—dissimilatory iron reduction; POC—particulate organic carbon. See Section 5.3 for discussion.

dependent Fe(II) oxidation (NDFO), the autotrophic/mixotrophic metabolism that couples the reduction of nitrate to nitrite with the oxidation of dFe(II) to pFe(III) (Straub et al., 1996; Kappler and Straub, 2005), or a catalyzed-oxidation of dFe(II) by nitrite (Picardal, 2012) (Fig. 11). Microbial catalysis of the oxidation of dFe(II) by nitrite has not been described, but this could also be a possibility.

5.4. A conceptual model for the redox cycling of Fe in oxygen minimum zones

The relative rates of supply and loss determine the concentration of dFe(II) in the ODZ. DFe(II) is supplied primarily from isopycnal transport from sedimentary DIR, augmented by some dFe(II) production from remineralization or Fe(II)-containing particulate organic matter by denitrifiers. DFe(II) is oxidized back to pFe(III) by nitrate or nitrite with the aid of microbial or abiotic catalysts, respectively. If the microbial abundance or concentrations of chemical catalysts are not high enough, it is likely that the oxidation of Fe(II) by nitrite proceeds too slowly to be important. Our data show that there is a higher Fe

oxidation rate closer to the margin: the most rapid decrease in dFe(II) with distance occurs in the stations closest to shore (Fig. 9). Indeed, while nitrite explains almost 50% of the variance of pFe in our dataset (Fig. 10), scatter about the best-fit line indicates that other factors are also important for the formation of pFe, which could include the availability of catalysts. We therefore hypothesize that there is a greater availability of catalysts (mineral or microbial) and thus higher oxidation rate near the margin. The greater availability of catalysts nearshore provides an effective Fe trap that removes most of the dFe(II).

If the offshore transport of dFe(II) is faster than the loss rate of dFe(II) by oxidation, as would occur if the availability of microbial or mineral surface catalysts decreased, this Fe trap may be leaky and lead to dFe(II) export into the interior, as we observed to Station 11. The presence of Ra-228, which has a half-life of 5.74 y, in the upper ODZ as far as Station 11 (Sanial et al., 2017), suggests a timescale of transport to Station 11 of months to years. This means that the timescale of oxidation of dFe(II), especially away from the margin, must be slow enough for dFe(II) to survive this transport.

6. CONCLUSIONS

Our direct observations of dissolved and particulate Fe in the Peruvian ODZ water column suggest that an effective redox trap prevents remobilized Fe from the shelf from being transported into the ocean interior, even when neither oxygen nor sulfide is present. Distributions of Fe species and hydrographic parameters point to nitrate or nitrite as a key player in the oxidation of dFe(II) in the absence of oxygen, but this oxidation requires the presence of microbial or abiotic catalysts, such as mineral surfaces from sediment resuspension. Certain environmental conditions may lead to a lack of appropriate catalysts, which may allow leaks in the redox trap so that dissolved Fe can penetrate further into the ocean interior. Future work should be directed to assessing the rates of microbial and abiotic anaerobic Fe oxidation in conditions more appropriate to oceanic ODZs.

ACKNOWLEDGEMENTS

This research was supported by NSF grant OCE-1518110 to P. J.L., OCE-1260273 to J.W.M., OCE-1232986 to B.M.T., and OCE-1233502 to K.W. Bruland, which funded C.P.T. The Advanced Light Source is supported by the Director, Office of Science, Office of Basic Energy Sciences, of the U.S. Department of Energy under Contract No. DE-AC02-05CH11231. Many thanks to Daniel Ohnemus, Sarah Nicholas, and the pump team for collecting and processing the particulate samples; the GEOTRACES sampling team (“supertechns”) for collecting the dissolved samples; the crew and chief scientists of the TG Thompson during the GEOTRACES EPZT cruise; the Scripps ODF team for hydrographic measurements and data management; Sarah Nicholas, Colleen Hoffman, Yang Xiang, Kira Mizell, Sirine Fakra, and Josep Roque-Rosell for assistance at beamline 10.3.2 at the Advanced Light Source; Carl Lamborg, Greg Cutter, and Virginie Sanial for discussions; four anonymous reviewers for comments that substantially improved the manuscript.

APPENDIX A. SUPPLEMENTARY MATERIAL

Supplementary data associated with this article can be found, in the online version, at <http://dx.doi.org/10.1016/j.gca.2017.05.019>.

REFERENCES

- Bange H. W., Naqvi S. W. A. and Codispoti L. A. (2005) The nitrogen cycle in the Arabian Sea. *Prog. Oceanogr.* **65**, 145–158.
- Benz M., Brune A. and Schink B. (1998) Anaerobic and aerobic oxidation of ferrous iron at neutral pH by chemoheterotrophic nitrate-reducing bacteria. *Arch. Microbiol.* **169**, 159–165.
- Biller D. V. and Bruland K. W. (2012) Analysis of Mn, Fe Co, Ni, Cu, Zn, Cd, and Pb in seawater using the Nobias-chelate PA1 resin and magnetic sector inductively coupled plasma mass spectrometry (ICP-MS). *Mar. Chem.* **130–131**, 12–20.
- Bishop J. K. B., Lam P. J. and Wood T. J. (2012) Getting good particles: accurate sampling of particles by large volume in-situ filtration. *Limnol. Oceanogr. Methods* **10**, 681–710.
- Bopp, L., Le Quere, C., Heimann, M., Manning, A.C., Monfray, P., 2002. Climate-induced oceanic oxygen fluxes: Implications for the contemporary carbon budget. *Global Biogeochem. Cycles* **16**.
- Boyd P. W., Jickells T., Law C. S., Blain S., Boyle E. A., Buesseler K. O., Coale K. H., Cullen J. J., de Baar H. J. W., Follows M., Harvey M., Lancelot C., Levasseur M., Owens N. P. J., Pollard R., Rivkin R. B., Sarmiento J., Schoemann V., Smetacek V., Takeda S., Tsuda A., Turner S. and Watson A. J. (2007) Mesoscale iron enrichment experiments 1993–2005: synthesis and future directions. *Science* **315**, 612–617.
- Bruchert V., Jorgensen B. B., Neumann K., Riechmann D., Schlosser M. and Schulz H. (2003) Regulation of bacterial sulfate reduction and hydrogen sulfide fluxes in the central Namibian coastal upwelling zone. *Geochim. Cosmochim. Acta* **67**, 4505–4518.
- Bruland K. W., Rue E. L., Smith G. J. and DiTullio G. R. (2005) Iron, macronutrients and diatom blooms in the Peru upwelling regime: brown and blue waters of Peru. *Mar. Chem.* **93**, 81–103.
- Buchwald C., Grabb K., Hansel C. M. and Wankel S. D. (2016) Constraining the role of iron in environmental nitrogen transformations: dual stable isotope systematics of abiotic NO₂-reduction by Fe(II) and its production of N₂O. *Geochim. Cosmochim. Acta* **186**, 1–12.
- Buresh R. J. and Moraghan J. T. (1976) Chemical reduction of nitrate by ferrous iron. *J. Environ. Qual.* **5**, 320–325.
- Canfield D. E., Stewart F. J., Thamdrup B., De Brabandere L., Dalsgaard T., Delong E. F., Revsbech N. P. and Ulloa O. (2010) A cryptic sulfur cycle in oxygen-minimum-zone waters off the Chilean coast. *Science* **330**, 1375–1378.
- Carlson L. and Schwertmann U. (1980) Natural occurrence of ferrixyhyte (d-FeOOH). *Clays Clay Miner* **28**, 272–280.
- Chelton D. B., Schlax M. G. and Samelson R. M. (2011) Global observations of nonlinear mesoscale eddies. *Prog. Oceanogr.* **91**, 167–216.
- Chever F., Rouxel O. J., Croot P. L., Ponzevera E., Wuttig K. and Auro M. (2015) Total dissolvable and dissolved iron isotopes in the water column of the Peru upwelling regime. *Geochim. Cosmochim. Acta* **162**, 66–82.
- Chukhrov F. V., Zvyagin B. B., Gorshkov A. I., Yermilova L. P., Korovushkin V. V., Rudnitskaya Y. S. and Yakubovskaya N. Y. (1977) Ferrixyhyte, a new modification of FeOOH. *Int. Geol. Rev.* **19**, 873–890.
- Coby A. J. and Picardal F. W. (2005) Inhibition of NO₃⁻ and NO₂⁻ reduction by microbial Fe(III) reduction: evidence of a reaction between NO₂⁻ and cell surface-bound Fe³⁺. *Appl. Environ. Microbiol.* **71**, 5267–5274.
- Cornell R. M. and Schwertmann U. (2003) *The Iron Oxides: Structure, Properties, Reactions, Occurrences and Uses*. John Wiley & Sons.
- Croot P. L. and Laan P. (2002) Continuous shipboard determination of Fe(II) in polar waters using flow injection analysis with chemiluminescence detection. *Anal. Chim. Acta* **466**, 261–273.
- Cutter G. A. and Bruland K. W. (2012) Rapid and noncontaminating sampling system for trace elements in global ocean surveys. *Limnol. Oceanogr.: Methods* **10**, 425–436.
- Cutter G. A., Andersson P., Codispoti L., Croot P., Francois F., Lohan M. C., Obata H., Rutgers van der Loeff M. (2014) Sampling and Sample-handling Protocols for GEOTRACES Cruises, v2.0. <<http://geotraces.org/images/stories/documents/intercalibration/Cookbook.pdf>>.
- Cutter G. A., Nielsdottir M. C. (2014). Characterizing redox conditions in oxygen-deficient waters off Peru during the 2013 US GEOTRACES Zonal Transect, Abstract OS22B-06, 2014 Fall Meeting, AGU. AGU, San Francisco, CA, pp. OS22B-06.
- Dalsgaard T., Thamdrup B., Farias L. and Revsbech N. P. (2012) Anammox and denitrification in the oxygen minimum zone of the eastern South Pacific. *Limnol. Oceanogr.* **57**, 1331–1346.
- Dugdale R. C., Goering J. J., Barber R. T., Smith R. L. and Packard T. T. (1977) Denitrification and hydrogen-sulfide in Peru upwelling region during 1976. *Deep-Sea Res.* **24**, 601–608.

- Hansen H. C. B., Borggaard O. K. and Sorensen J. (1994) Evaluation of the free energy of formation of Fe(II)-Fe(III) hydroxide-sulphate (green rust) and its reduction of nitrite. *Geochim. Cosmochim. Acta* **58**, 2599–2608.
- Heller M. I., Moffett J. W., Tiano L., Revsbech N. P. (2013). Distribution of Fe(II), H_2O_2 and O_2 in the Mexican OMZ: potential biogeochemical processes causing their occurrence. Abstracts of Papers of the American Chemical Society, p. 245.
- Hong H. S. and Kester D. R. (1986) Redox state of iron in offshore waters of Peru. *Limnol. Oceanogr.* **31**, 512–524.
- Hopkinson B. M. and Barbeau K. A. (2007) Organic and redox speciation of iron in the eastern tropical North Pacific suboxic zone. *Mar. Chem.* **106**, 2–17.
- Johnson K. S., Coale K. H., Elrod V. A. and Neil W. T. (1994) Iron photochemistry in seawater from the equatorial Pacific. *Mar. Chem.* **46**, 319–334.
- Jones A. M., Griffin P. J., Collins R. N. and Waite T. D. (2014) Ferrous iron oxidation under acidic conditions – the effect of ferric oxide surfaces. *Geochim. Cosmochim. Acta* **145**, 1–12.
- Jones L. C., Peters B., Lezama Pacheco J. S., Casciotti K. L. and Fendorf S. (2015) Stable isotopes and iron oxide mineral products as markers of chemodenitrification. *Environ. Sci. Technol.* **49**, 3444–3452.
- Kappler A. and Straub K. L. (2005) Geomicrobiological cycling of iron. *Rev. Mineral. Geochem.* **59**, 85–108.
- Kessler W. S. (2006) The circulation of the eastern tropical Pacific: a review. *Prog. Oceanogr.* **69**, 181–217.
- King D. W., Lounsbury H. A. and Millero F. J. (1995) Rates and mechanism of Fe(II) oxidation at nanomolar total iron concentrations. *Environ. Sci. Technol.* **29**, 818–824.
- Klueglein N., Picardal F., Zedda M., Zwiener C. and Kappler A. (2015) Oxidation of Fe(II)-EDTA by nitrite and by two nitrate-reducing Fe(II)-oxidizing Acidovorax strains. *Geobiology* **13**, 198–207.
- Kondo Y., Moffett J. W., Obata H. and Nishioka J. (2010) Fe(II) distribution in the Arabian Sea oxygen minimum zone and Western tropical Indian Ocean from GEOTRACES KH-09-5. *Geochim. Cosmochim. Acta* **74**, A528. A528.
- Kondo Y. and Moffett J. W. (2013) Dissolved Fe(II) in the Arabian Sea oxygen minimum zone and western tropical Indian Ocean during the inter-monsoon period. *Deep Sea Res. Part I* **73**, 73–83.
- Kondo Y. and Moffett J. W. (2015) Iron redox cycling and subsurface offshore transport in the eastern tropical South Pacific oxygen minimum zone. *Mar. Chem.* **168**, 95–103.
- Kopf S. H., Henny C. and Newman D. K. (2013) Ligand-enhanced abiotic iron oxidation and the effects of chemical versus biological iron cycling in anoxic environments. *Environ. Sci. Technol.* **47**, 2602–2611.
- Lam P. J., Ohnemus D. C. and Marcus M. A. (2012) The speciation of marine particulate iron adjacent to active and passive continental margins. *Geochim. Cosmochim. Acta* **80**, 108–124.
- Landing W. M. and Bruland K. W. (1987) The contrasting biogeochemistry of iron and manganese in the Pacific-Ocean. *Geochim. Cosmochim. Acta* **51**, 29–43.
- Lyons T. W. and Severmann S. (2006) A critical look at iron paleoredox proxies: new insights from modern euxinic marine basins. *Geochim. Cosmochim. Acta* **70**, 5698–5722.
- Marcus M. A., MacDowell A. A., Celestre R., Manceau A., Miller T., Padmore H. A. and Sublett R. E. (2004) Beamline 10.3.2 at ALS: a hard X-ray microprobe for environmental and materials sciences. *J. Synchrotron Radiat.* **11**, 239–247.
- Marcus M. A. (2010) X-ray photon-in/photon-out methods for chemical imaging. *Trends Anal. Chem.* **29**, 508–517.
- Marsay C. M. and John S. G. (2014) *Distribution and Isotopic Signature of “Ligand-leachable” Particulate Iron in the Eastern Tropical South Pacific: Initial Results from US GEOTRACES EPZT, Abstract OS23E-1268, 2014 Fall Meeting. AGU, AGU, San Francisco, CA.*
- Millero F. J., Sotolongo S. and Izaguirre M. (1987) The oxidation-kinetics of Fe(II) in seawater. *Geochim. Cosmochim. Acta* **51**, 793–801.
- Millero F. J. and Sotolongo S. (1989) The oxidation of Fe(II) with H_2O_2 in seawater. *Geochim. Cosmochim. Acta* **53**, 1867–1873.
- Moffett J. W., Goepfert T. J. and Naqvi S. W. A. (2007) Reduced iron associated with secondary nitrite maxima in the Arabian Sea. *Deep Sea Res. Part I* **54**, 1341–1349.
- Moffett J. W. and Zika R. G. (1987) Reaction kinetics of hydrogen peroxide with copper and iron in seawater. *Environ. Sci. Technol.* **21**, 804–810.
- Moraghan J. T. and Buresh R. J. (1977) Chemical reduction of nitrite and nitrous oxide by ferrous iron. *Soil Sci. Soc. Am. J.* **41**, 47–50.
- Noffke A., Hensen C., Sommer S., Scholz F., Bohlen L., Mosch T., Graco M. and Wallmann K. (2012) Benthic iron and phosphorus fluxes across the Peruvian oxygen minimum zone. *Limnol. Oceanogr.* **57**, 851–867.
- Ohnemus D. C., Auro M. E., Sherrell R. M., Lagerstrom M., Morton P. L., Twining B. S., Rauschenberg S. and Lam P. J. (2014) Laboratory intercomparison of marine particulate digestions including Piranha: a novel chemical method for dissolution of polyethersulfone filters. *Limnol. Oceanogr. – Methods* **12**, 530–547.
- Ohnemus D. C. and Lam P. J. (2015) Cycling of lithogenic marine particles in the US GEOTRACES North Atlantic transect. *Deep Sea Res. Part II* **116**, 283–302.
- Ohnemus D. C., Rauschenberg S., Cutter G. A., Fitzsimmons J. N., Sherrell R. M. and Twining B. S. (2016) Elevated trace metal content of prokaryotic communities associated with marine oxygen deficient zones. *Limnol. Oceanogr.*
- Parker C. E., Brown M. T. and Bruland K. W. (2016) Scandium in the open ocean: a comparison with other group 3 trivalent metals. *Geophys. Res. Lett.* **43**, 2758–2764.
- Picardal F. (2012). Abiotic and microbial interactions during anaerobic transformations of Fe(II) and NOX. *Front. Microbiol.* **3**.
- Raiswell R. and Canfield D. E. (2012) The iron biogeochemical cycle past and present. *Geochem. Perspect.* **1**, 1–220.
- Revsbech N. P., Larsen L. H., Gundersen J., Dalsgaard T., Ulloa O. and Thamdrup B. (2009) Determination of ultra-low oxygen concentrations in oxygen minimum zones by the STOX sensor. *Limnol. Oceanogr. – Methods* **7**, 371–381.
- Rose A. L. and Waite T. D. (2001) Chemiluminescence of luminol in the presence of iron (II) and oxygen: oxidation mechanism and implications for its analytical use. *Anal. Chem.* **73**, 5909–5920.
- Roy E. G., Wells M. L. and King D. W. (2008) Persistence of iron (II) in surface waters of the western subarctic Pacific. *Limnol. Oceanogr.* **53**, 89–98.
- Rue E. L., Smith G. J., Cutter G. A. and Bruland K. W. (1997) The response of trace element redox couples to suboxic conditions in the water column. *Deep-Sea Res. Part I – Oceanogr. Res. Papers* **44**, 113–134.
- Sanial V., Kipp L. E., Henderson P., van Beek P., Reyss J. L., Hammond D. E., Hawco N. J., Saito M. A., Resing J. A., Sedwick P., Moore W. S. and Charette M. A. (2017) Radium-228 as a tracer of dissolved trace element inputs from the Peruvian continental margin. *Mar. Chem.* <http://dx.doi.org/10.1016/j.marchem.2017.05.008>.

- Scholz F., McManus J., Mix A. C., Hensen C. and Schneider R. R. (2014a) The impact of ocean deoxygenation on iron release from continental margin sediments. *Nat. Geosci.* **7**, 433–437.
- Scholz F., Severmann S., McManus J. and Hensen C. (2014b) Beyond the Black Sea paradigm: the sedimentary fingerprint of an open-marine iron shuttle. *Geochim. Cosmochim. Acta* **127**, 368–380.
- Scholz F., Löscher C. R., Fiskal A., Sommer S., Hensen C., Lomnitz U., Wuttig K., Göttlicher J., Kossel E., Steininger R. and Canfield D. E. (2016) Nitrate-dependent iron oxidation limits iron transport in anoxic ocean regions. *Earth Planet. Sci. Lett.* **454**, 272–281.
- Schwertmann U. and Thalmann H. (1976) The influence of [Fe(II)], [Si], and pH on the formation of lepidocrocite and ferrihydrite during oxidation of aqueous FeCl₂ solutions. *Clay Miner.* **11**, 189–200.
- Schwertmann U. and Fechter H. (1994) The formation of green rust and its transformation to Lepidocrocite. *Clay Miner.* **29**, 87–92.
- Senn A.-C., Kaegi R., Hug S. J., Hering J. G., Mangold S. and Voegelin A. (2015) Composition and structure of Fe(III)-precipitates formed by Fe(II) oxidation in water at near-neutral pH: interdependent effects of phosphate, silicate and Ca. *Geochim. Cosmochim. Acta* **162**, 220–246.
- Sorensen J. and Thorling L. (1991) Stimulation by lepidocrocite (γ -FeOOH) of Fe(II)-dependent nitrite reduction. *Geochim. Cosmochim. Acta* **55**, 1289–1294.
- Stramma L., Johnson G. C., Sprintall J. and Mohrholz V. (2008) Expanding oxygen-minimum zones in the tropical oceans. *Science* **320**, 655–658.
- Straub K. L., Benz M., Schink B. and Widdel F. (1996) Anaerobic, nitrate-dependent microbial oxidation of ferrous iron. *Appl. Environ. Microbiol.* **62**, 1458–1460.
- Tai Y.-L. and Dempsey B. A. (2009) Nitrite reduction with hydrous ferric oxide and Fe(II): stoichiometry, rate, and mechanism. *Water Res.* **43**, 546–552.
- Tiano L., Garcia-Robledo E., Dalsgaard T., Devol A. H., Ward B. B., Ulloa O., Canfield D. E. and Peter Revsbech N. (2014) Oxygen distribution and aerobic respiration in the north and south eastern tropical Pacific oxygen minimum zones. *Deep Sea Res. Part I* **94**, 173–183.
- Toner B. M., German C. R., Dick G. J. and Breier J. A. (2016) Deciphering the complex chemistry of deep-ocean particles using complementary synchrotron X-ray microscope and microprobe instruments. *Acc. Chem. Res.* **49**, 128–137.
- Ulloa O., Canfield D. E., DeLong E. F., Letelier R. M. and Stewart F. J. (2012) Microbial oceanography of anoxic oxygen minimum zones. *Proc. Natl. Acad. Sci.* **109**, 15996–16003.
- Van Cleemput O. and Baert L. (1983) Nitrite stability influenced by iron compounds. *Soil Biol. Biochem.* **15**, 137–140.
- Vedamati J., Goepfert T. and Moffett J. W. (2014) Iron speciation in the eastern tropical South Pacific oxygen minimum zone off Peru. *Limnol. Oceanogr.* **95**, 1945–1957.
- Voegelin A., Kaegi R., Frommer J., Vantelon D. and Hug S. J. (2010) Effect of phosphate, silicate, and Ca on Fe(III)-precipitates formed in aerated Fe(II)- and As(III)-containing water studied by X-ray absorption spectroscopy. *Geochim. Cosmochim. Acta* **74**, 164–186.
- Voegelin A., Senn A.-C., Kaegi R., Hug S. J. and Mangold S. (2013) Dynamic Fe-precipitate formation induced by Fe(II) oxidation in aerated phosphate-containing water. *Geochim. Cosmochim. Acta* **117**, 216–231.
- von der Heyden B. P., Hauser E. J., Mishra B., Martinez G. A., Bowie A. R., Tyliczszak T., Mtshali T. N., Roychoudhury A. N. and Myneni S. C. B. (2014) Ubiquitous presence of Fe(II) in aquatic colloids and its association with organic carbon. *Environ. Sci. Technol. Lett.*
- Weber K. A., Pollock J., Cole K. A., O'Connor S. M., Achenbach L. A. and Coates J. D. (2006) Anaerobic nitrate-dependent iron (II) bio-oxidation by a novel lithoautotrophic betaproteobacterium, strain 2002. *Appl. Environ. Microbiol.* **72**, 686–694.
- Wullstein L. H. and Gilmour C. M. (1966) Non-enzymatic formation of nitrogen gas. *Nature* **210**, 1150–1151.
- Zeng T., Arnold W. A. and Toner B. M. (2013) Microscale characterization of sulfur speciation in lake sediments. *Environ. Sci. Technol.* **47**, 1287–1296.
- Zika R. G., Saltzman E. S. and Cooper W. J. (1985) Hydrogen-peroxide concentrations in the Peru upwelling area. *Mar. Chem.* **17**, 265–275.

Associate Editor: Jack J. Middelburg